

CHAPTER 6

WASTEWATER TREATMENT AND DISCHARGE

Final Draft

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6 WASTEWATER TREATMENT AND DISCHARGE

Users are expected to go to Mapping Tables in Annex 1, before reading this chapter. This is required to correctly understand both the refinements made and how the elements in this chapter relate to the corresponding chapter in the 2006 IPCC Guidelines.

6.1 INTRODUCTION

Wastewater can be a source of methane (CH₄) when treated or disposed anaerobically or when dissolved CH₄ enters aerated treatment systems. It can also be a source of nitrous oxide (N₂O) emissions. Carbon dioxide (CO₂) emissions from wastewater are not considered in the *IPCC Guidelines* because these are generally derived from modern (biogenic) organic matter in human excreta or food waste and should not be included in national total emissions. Appendix 6A.1 presents a discussion of abiogenic (fossil) CO₂ emissions from wastewater treatment and discharge, where fossil organic carbon is present in wastewater or treatment sludge.

Wastewater originates from a variety of domestic, commercial and industrial sources and may be treated on site (uncollected), sewer to a centralized plant (collected) or disposed untreated nearby or via an outfall. Domestic wastewater is defined as wastewater from household water use, while industrial wastewater is from industrial practices only.¹ Treatment and discharge systems can sharply differ between countries. Also, treatment and discharge systems can differ for rural and urban users, and for urban high income and urban low-income users.

Sewers may be open or closed. In urban areas in developing countries and some developed countries, sewer systems may consist of networks of open canals, gutters, and ditches, which are referred to as open sewers. These systems are subject to heating from the sun and the sewers may be stagnant allowing for anaerobic conditions to emit CH₄ (Doorn *et al.* 1997). In most developed countries and in high-income urban areas in other countries, sewers are usually closed and underground. Wastewater in closed underground sewers likely generate CH₄, but there are insufficient data available to quantify the emissions from these collection systems. However, research shows that significant amounts of CH₄ can be formed within closed sewer collection systems and enters centralised wastewater treatment plants (WWTPs) as dissolved CH₄ in the wastewater, where it is then emitted during treatment.

The most common wastewater treatment methods in developed countries are centralized aerobic wastewater treatment plants and lagoons for both domestic and industrial wastewater. To avoid high discharge fees or to meet regulatory standards, many large industrial facilities pre-treat their wastewater before releasing it into the sewage system. Domestic wastewater may also be treated in on-site septic systems. These are advanced systems that may treat wastewater from one or several households. They consist of an anaerobic underground tank and a drainage field for the treatment of effluent from the tank. Some developed countries continue to dispose of untreated domestic wastewater via an outfall or pipeline into a water body, such as the ocean.

The degree of wastewater treatment varies in most developing countries. In some cases industrial wastewater is discharged directly into bodies of water, while major industrial facilities may have comprehensive in-plant treatment. Domestic wastewater is treated in centralized plants, pit latrines, septic systems or disposed of in unmanaged lagoons or waterways, via open or closed sewers. In some coastal cities domestic wastewater is discharged directly into the ocean. Pit latrines are lined or unlined holes of up to several meters deep, which may be fitted with a toilet for convenience.

Centralized wastewater treatment methods can be classified as primary, secondary, and tertiary treatment. In primary treatment, physical barriers remove larger solids from the wastewater. Remaining particulates are then allowed to settle. Secondary treatment consists of a combination of biological processes that promote biodegradation by micro-organisms. These may include aerobic stabilisation ponds, trickling filters, and activated sludge processes, as well as anaerobic reactors and lagoons. Tertiary treatment processes are used to further purify the wastewater of pathogens, contaminants, and remaining nutrients such as nitrogen and phosphorus compounds. This is achieved using one or a combination of processes that can include maturation/polishing ponds, biological processes, advanced filtration, carbon adsorption, ion exchange, and disinfection.

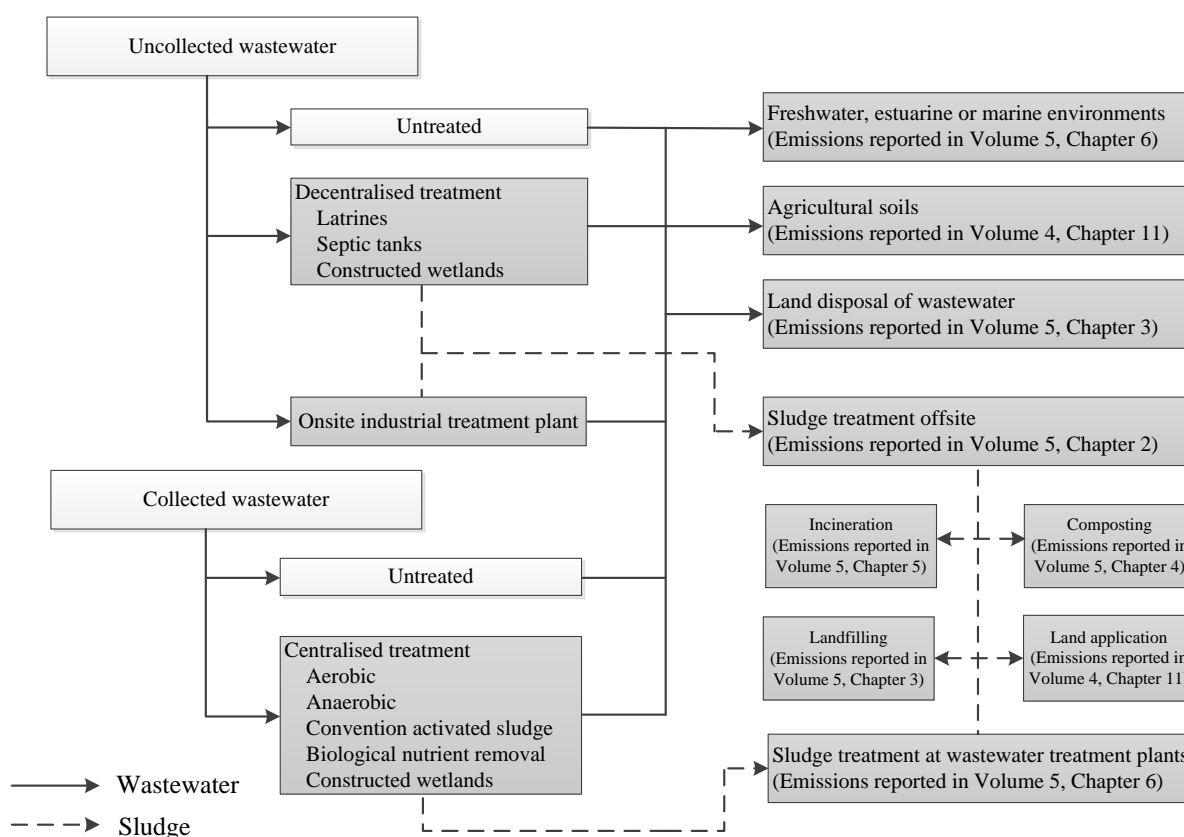
Sludge is produced in all of the primary, secondary and tertiary stages of treatment. Sludge that is produced in primary treatment consists of solids that are removed from the wastewater and is not accounted for in this category. Sludge produced in secondary and tertiary treatment results from biological growth in the biomass, as well as the

¹ Because the methodology is on a per person basis, emissions from commercial wastewater are estimated as part of domestic wastewater. To avoid confusion, the term municipal wastewater is not used in this text. Municipal wastewater is a mix of household, commercial and non-hazardous industrial wastewater, treated at wastewater treatment plants.

collection of small particles. This sludge must be treated further before it can be safely disposed of. Methods of sludge treatment include aerobic and anaerobic stabilisation (digestion), conditioning, centrifugation, composting, and drying. Land disposal, composting, and incineration of sludge is considered in Volume 5, Section 2.3.2 in Chapter 2, Waste Generation, Composition, and Management Data, Section 3.2 in Chapter 3, Solid Waste Disposal, Section 4.1 in Chapter 4, Biological Treatment and Disposal, and Chapter 5, Incineration and Open Burning of Waste, respectively. Some sludge is incinerated before land disposal. N₂O emissions from sludge and wastewater spread on agricultural land are considered in Section 11.2, N₂O emissions from managed soils, in Chapter 11, N₂O Emissions from Managed Soils, and CO₂ Emissions from Lime and Urea Application, in Volume 4 of the Agriculture, Forestry, and Other Land Use (AFOLU) Sector.

Figure 6.1 from the 2006 IPCC Guidelines has been updated and simplified to show the different pathways for wastewater treatment and discharge, for wastewater that is not collected and for wastewater that is collected. The figure specifies whether the discharge or end use of the wastewater is reported in this chapter. Figure 6.1 (Updated) also shows sludge treatment pathways in grey and clarifies whether the emissions are reported in this chapter. Table 6.1 has been updated to reflect the main wastewater treatment and discharge systems in developed and developing countries and their potential to emit CH₄ and N₂O.

Figure 6.1 (Updated) Wastewater treatment systems and discharge pathways



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TABLE 6.1 (UPDATED)				
CH ₄ AND N ₂ O EMISSION POTENTIALS FOR WASTEWATER AND SLUDGE TREATMENT AND DISCHARGE SYSTEMS				
Types of treatment and disposal			CH ₄ and N ₂ O emission potentials	
Discharge from Collected or Uncollected Systems	Untreated or Treated Systems	Freshwater, estuarine, or marine discharge		While modulated by oxygen status, CH ₄ is generated in a range of freshwater and estuarine environments. Among them, stagnant or oxygen deficient environments are probable sources of N ₂ O.
		Non-aquatic environment (soils)		While modulated by climatic conditions, CH ₄ is generated in soil environment. N ₂ O is generated.
Collected	Untreated	Sewers (closed and underground)		Likely source of CH ₄ /N ₂ O. However, insufficient data exist to quantify emission factors that address the variation in sewer type and operational conditions.
		Sewers (open)		Stagnant, overloaded open collection sewers or ditches/canals are likely significant sources of CH ₄ .
	Treated	Aerobic treatment	Centralised aerobic wastewater treatment plants	May produce limited CH ₄ from anaerobic pockets. May also liberate CH ₄ generated in upstream sewer networks during turbulent and/or aerobic treatment processes. Poorly designed or managed aerobic treatment systems produce higher CH ₄ due to reduced removal of organics in sludge during primary treatment. Plants with nutrient removal processes are sources of CH ₄ and N ₂ O.
			Aerobic shallow ponds	Unlikely source of CH ₄ /N ₂ O. Poorly designed or managed aerobic systems produce CH ₄ .
		Anaerobic treatment	Anaerobic lagoons	May be a significant source of CH ₄ . Insignificant source of N ₂ O.
			Facultative lagoons ²	Source of CH ₄ from anaerobic layer.
			Constructed wetlands	May be source of CH ₄ and N ₂ O. See <i>2013 Supplement to the 2006 IPCC Guidelines for National Greenhouse Gas Inventories: Wetlands</i> (IPCC 2014).
			Anaerobic reactors	May be a significant source of CH ₄ if emitted CH ₄ is not recovered or flared.
		Onsite sludge treatment ¹	Sludge anaerobic treatment in centralised aerobic wastewater treatment plant	Sludge may be a significant source of CH ₄ if emitted CH ₄ is not recovered or flared. In addition, sludge digestion and handling may be a source of fugitive CH ₄ from biogas recovery operations. See Chapter 4 for more details.
			Composting	May be significant source of CH ₄ and N ₂ O. See Chapter 4 for more details.

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TABLE 6.1 (UPDATED) (CONTINUED)				
CH ₄ AND N ₂ O EMISSION POTENTIALS FOR WASTEWATER AND SLUDGE TREATMENT AND DISCHARGE SYSTEMS				
Types of treatment and disposal				CH ₄ and N ₂ O emission potentials
			Incineration and open burning	Source of CH ₄ and N ₂ O. See Chapter 5 for more details.
Uncollected	Septic tanks (without dispersion field)			Source of CH ₄ . Frequent solids removal reduces CH ₄ production.
	Septic system (including a septic tank and a soil dispersal system)			Source of CH ₄ (tank) and N ₂ O (soil dispersal system) Frequent solids removal reduces CH ₄ production.
	Open pits/Latrines			Pits/latrines are likely to produce CH ₄ when temperature and retention time are favourable.
¹ For onsite sludge treatment, see Chapters 4 and 5 for emissions methodology, but note that emissions for onsite systems should be reported under the Wastewater Treatment and Discharge category.				
² Facultative organisms can function in the presence or absence of molecular oxygen. In a facultative lagoon, the layer of water near the surface contains dissolved oxygen due to atmospheric reaeration and algal respiration, a condition that supports aerobic and facultative organisms. The bottom layer of the lagoon includes sludge deposits and supports anaerobic organisms. The intermediate anoxic layer—the facultative zone—ranges from aerobic near the top to anaerobic at the bottom (US EPA 2002b).				

6.1.1 Centralised treatment systems

Centralised wastewater treatment systems may include a variety of treatment processes spanning the hierarchy of treatment levels. Primary treatment involves mechanical processes such as screening, grit removal and sedimentation. Secondary treatment involves biological processes in which microorganisms convert inorganic and organic nutrients into settleable solids which can be removed by sedimentation and degradation metabolites which are transferred to the atmosphere. Secondary biological treatment processes may be augmented with chemical disinfection. Commonly used secondary treatment processes include: activated sludge; trickling filters; and lagoons. Tertiary treatment involves the use of advanced biological nutrient removal processes (nitrification–denitrification), and other physical and chemical processes. Treatment processes can range from simple technologies such as lagooning and wetlands, to more technologically advanced treatment technology designed for stringent nutrient removal. Wetlands can be constructed or semi-natural systems and may be used as the primary method of wastewater treatment, or as a polishing treatment step following settling and biological treatment. Constructed wetlands develop natural processes that involve vegetation, soil, and associated microbial assemblages to trap and treat incoming contaminants (IPCC 2014).

Sewer collection systems provide an environment conducive to the formation of CH₄, which can be substantial depending on the configuration and operation of the collection system (Guisasola *et al.* 2008). Recent research has shown that at least a portion of CH₄ formed with the collection system enters the centralised system where it contributes to CH₄ emissions from the treatment system (Foley *et al.* 2015). Although there are insufficient data to quantify emissions directly from the sewer collection system, the emission factors presented for centralised treatment plants account for dissolved methane entering the treatment systems.

Soluble organic matter is generally removed using biological processes in which microorganisms consume the organic matter for maintenance and growth. The resulting biomass (sludge) is removed from the effluent prior to discharge to receiving environments. Microorganisms can biodegrade soluble organic material in wastewater under aerobic and/or anaerobic conditions, where the latter condition produces CH₄. During collection and treatment, wastewater may be accidentally or deliberately managed under anaerobic conditions. In addition, the sludge may be further biodegraded under aerobic or anaerobic conditions, such as with anaerobic digestion.

The generation of N₂O may also result from the treatment of domestic wastewater during both nitrification and denitrification of the nitrogen (N) present, usually in the form of urea, proteins, and ammonia. Ammonia N is converted to nitrate (NO₃[−]) through the aerobic process of nitrification. Denitrification occurs under anoxic/anaerobic conditions, whereby aerobic or facultative organisms reduce oxidized forms of nitrogen (e.g., nitrite, nitrate) in the absence of free oxygen to produce nitrogen gas (N₂). N₂O is an intermediate product of both nitrification and denitrification processes. No matter where N₂O is formed it is usually stripped to the air in aerated parts of the treatment process.

A common example of a centralised treatment system configuration is an activated sludge wastewater treatment system. Following grit removal and primary treatment for solids removal, wastewater is sent to an activated sludge reactor for carbonaceous biochemical oxygen demand removal, followed by secondary clarification where solids are allowed to settle from the wastewater. Clarified effluent may be disinfected prior to discharge. Secondary

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clarifier sludge is pumped out from the bottom of the clarifier. Of this sludge, a portion is sent back to the activated sludge treatment process (return activated sludge) and the remainder (waste activated sludge) is combined with primary treatment sludge before being sent to sludge handling (such as gravity thickening). The thickened sludge may also be processed onsite in an anaerobic digester followed by further dewatering, such as by centrifuge. Recovered wastewater from thickening and dewatering operations, such as filtrate from the gravity thickener or centrate from the centrifuge) are returned to the influent stream at the headworks to the wastewater treatment system.

Although the CH₄ emissions from centralised aerobic plants may be small, they are not zero and may be substantial for some plants receiving sewage from expansive sewer networks (Willis 2017). In addition, anaerobic conditions can form within the treatment system increasing the potential for CH₄ generation from an otherwise aerobic system.

CH₄ generated in the sewer system enters the treatment plant as dissolved methane and is emitted during turbulent/aerated treatment steps (Daelman *et al.* 2012). These refinements introduce a new MCF associated with these systems, as well as a revised N₂O emission factor for centralised wastewater treatment systems.

Sometimes conditions form that result in the wastewater treatment system becoming hydraulically or organically overloaded, which results in degraded performance of the plant. In these cases, organic matter and nutrients that would normally be removed by the system instead pass through and are discharged to the aquatic environment. The 2006 IPCC Guidelines previously provided different emission factors for “well managed” centralised aerobic treatment systems versus “not well managed” centralised aerobic treatment systems. However, these refinements remove those factors and instead provide one factor for these types of systems. In the case of overloaded systems, the amount of organics removed by the system in sludge (S) should reflect the performance achieved by the system.

6.1.2 Decentralised treatment systems of domestic wastewater (onsite sanitation)

Depending on national circumstances, domestic wastewater not connected to a centralised wastewater treatment plant may be directly discharged in natural aquatic environments (rivers, lakes, oceans, etc.) or treated using onsite sanitation. The most common onsite treatment systems are holding tanks, septic systems and latrines.

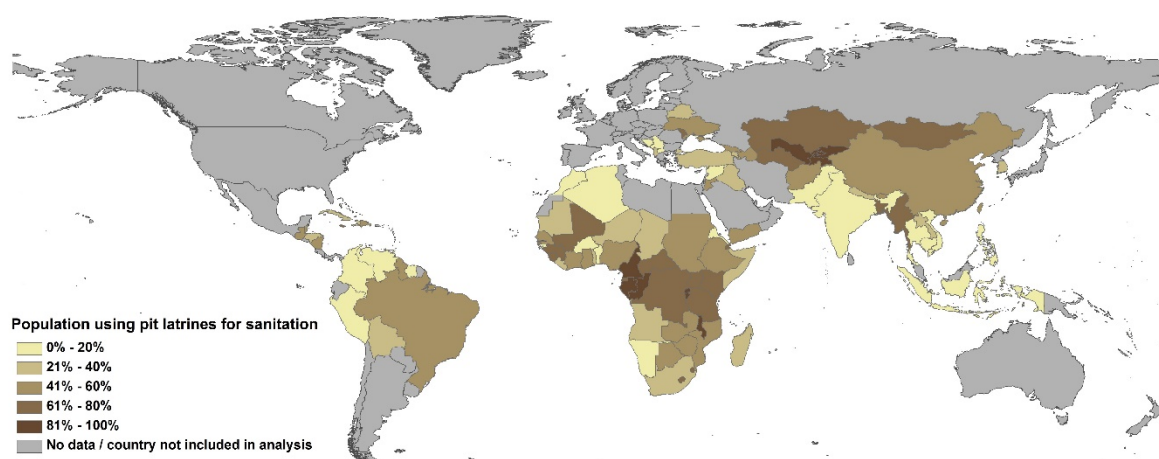
A septic system is usually composed of a septic tank, generally buried in the ground, and a soil dispersal system. Solids and dense materials contained in the incoming wastewater (influent) settle in the septic tanks as sludge. Floatable material (scum) is also retained in the tank. The sludge settled on the bottom of the tank undergoes anaerobic digestion. Partially treated water is discharged in the dispersal system. The liquid fraction remains in the tank for only a short period, with the hydraulic retention time (HRT) varying from 24 to 72 hours depending on tank volume and hydraulic load. The solid fraction accumulates and remains in the tank for several years, during which time it degrades anaerobically. The solids retention time (SRT) depends on the sludge withdrawal frequency. The gas produced from anaerobic sludge digestion (mainly CH₄ and largely biogenic CO₂) rise to the liquid surface and are usually released through vents. Gases produced in the effluent dispersal system (mainly N₂O and biogenic CO₂) are released through the soil.

In some high-income countries, onsite aerated wastewater treatment systems are used and enable a more advanced level of treatment than septic tanks in reducing the load of organics and nutrients in domestic effluent. The process usually involves a first step of sedimentation and anaerobic digestion, a second step of aerobic treatment and last step of clarification and disinfection. The treated effluent is discharged into the environment via surface irrigation or infiltration through an absorption trench.

A latrine usually consists of a slab over a pit which may be two meters or more in depth. A wide range of configuration options exists for latrines (simple pit latrines, ventilated latrines, composting latrines, etc.) having in common that little (e.g., pour flush latrines) or no water is used to flush excreta into the pit. Pit latrines are utilised by more than 1.5 billion people throughout the world, especially in low-income countries (see new Figure 6.1a and new Annex 6A.1). Pit latrine gaseous emissions depend in part on local groundwater level. Anaerobic conditions favourable to CH₄ emissions occur when the water table is high and the organic waste in the pit is submerged.

In the absence of latrines, people resort to open defecation. Open defecation is not considered as a source of CH₄, as anaerobic conditions are considered unlikely.

Figure 6.1a (New) Percentage of low-income country populations using pit latrines as a primary sanitation facility (Graham & Polizzotto 2013)



6.1.3 Emissions from receiving waters

Dissolved CH₄ and N₂O that is generated in sewers, or present in untreated or primary treated discharges, has the potential to be released (Short *et al.* 2014; Short *et al.* 2017). A strong correlation between the condition of the aquatic environment and the generation of CH₄ and N₂O has been observed (e.g., Smith *et al.* 2017). Therefore, where wastewater is then discharged to aquatic environments with nutrient-impacted/eutrophic conditions (i.e., water bodies which are rich in nutrients and very productive in terms of aquatic animal and plant life), the additional organic matter in the discharged wastewater is expected to increase emissions. Many waterways are naturally eutrophic, while others have been altered by human impacts and are subject to eutrophication as a result. In the case of CH₄, most freshwaters are highly supersaturated with CH₄ irrespective of their trophic state, so this refinement does not distinguish between eutrophic and oligotrophic receiving waters, but it does distinguish between default waters (rivers and streams) and lakes and reservoirs where CH₄ emissions are higher.

6.1.4 Changes compared to 1996 Guidelines and Good Practice Guidance

No refinement.

6.1.5 Changes compared to 2006 Guidelines

The 2006 IPCC Guidelines included combined equations to estimate CH₄ emissions from wastewater and from sludge removed from the wastewater. However, in some cases, this combined equation caused confusion among inventory compilers when calculating CH₄ emissions from aerobic systems with anaerobic sludge digestion. In these cases, some compilers estimated zero CH₄ emissions from well operated wastewater treatment systems, and then subtracted emissions associated with sludge digestion operations without first estimating the CH₄ emissions from sludge treatment, resulting in negative emissions. In this refinement, we discuss the proper way to use the equation in such situations, and we present an update to provide guidance on the calculation of the organic component removed in sludge. Countries must estimate the amount (mass) of sludge they generate from wastewater treatment and default data are provided for a number of countries. With this additional guidance, the use of a default value of zero for sludge removal from aerobic treatment systems and septic systems is no longer applicable.

In addition, certain emission factors for CH₄ emissions from domestic and industrial wastewater treatment have been updated to reflect additional measurement data on emissions from septic systems and centralised WWTPs. Furthermore, the CH₄ emission factors for wastewater discharged to aquatic environments have been updated and a new emission factor for discharge to reservoirs, lakes, and estuaries is introduced. The calculation of CH₄ emissions from effluent discharged to aquatic systems has been updated to include the discharge of treated effluent and to reflect the removal of organics that occurs during treatment.

The 2019 Refinement also includes new guidance on how to estimate N₂O emissions from domestic and industrial wastewater and presents updated guidance to estimate N₂O emissions from centralised WWTPs. Furthermore, the

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N₂O emission factors for wastewater discharged to aquatic environments have been updated and the calculation of N₂O emissions from effluent discharged to aquatic systems has been updated to reflect the removal of nitrogen that occurs during treatment.

6.2 METHANE EMISSIONS FROM WASTEWATER

6.2.1 Methodological issues

Emissions are a function of the amount of organic waste generated and an emission factor that characterises the extent to which this waste generates CH₄.

Three tier methods for CH₄ from this category are summarised below:

The Tier 1 method applies default values for the emission factor and activity parameters. This method is considered *good practice* for countries with limited data.

The Tier 2 method follows the same method as Tier 1 but allows for incorporation of a country specific emission factor and country specific activity data. For example, a specific emission factor for a prominent treatment system based on field measurements could be incorporated under this method. The amount of sludge removed for incineration, landfills, and agricultural land should be taken into consideration.

For a country with good data and advanced methodologies, a country specific method could be applied as a Tier 3 method. For example, a more advanced country-specific method could be based on plant-specific data from large wastewater treatment facilities, using country-specific measurements of organics discharged to aquatic environments, or may draw on country-specific water quality data for aquatic environments receiving wastewater inputs.

Wastewater treatment facilities that receive wastewater from collection systems, particularly pressurized sewers and gravity-fed sewers that are closed, can liberate CH₄ in aerobic systems from dissolved CH₄ that enters the treatment system (Daelman *et al.* 2012; Short *et al.* 2017).

Wastewater treatment facilities can include anaerobic process steps. CH₄ generated at such facilities can be recovered and combusted in a flare or energy device. The amount of CH₄ that is flared or recovered for energy use should be subtracted from total emissions through the use of a separate CH₄ recovery parameter. The amount of CH₄ which is recovered or flared is expressed as R in updated Equation 6.1.

Note that only a few countries may have sludge removal data and CH₄ recovery data. The information on sludge generation has improved and statistical data on sludge can be found in databases of Eurostat and OECD. The lack of data for CH₄ recovery was taken into account in this refinement by recommending the use of the methodology in Section 4.1, Chapter 4 of Volume 5, *2006 IPCC Guidelines* which estimates emissions on the basis of total sludge anaerobically digested; thus, the information on R is not required in this calculation. Default sludge removal factors are provided in this refinement. The default for CH₄ recovery is zero. If a country selects to report CH₄ recovery, it is *good practice* to distinguish between flaring and CH₄ recovery for energy generation, which should be reported in the Energy Sector taking into account the avoidance of double counting emissions from flaring and energy used.

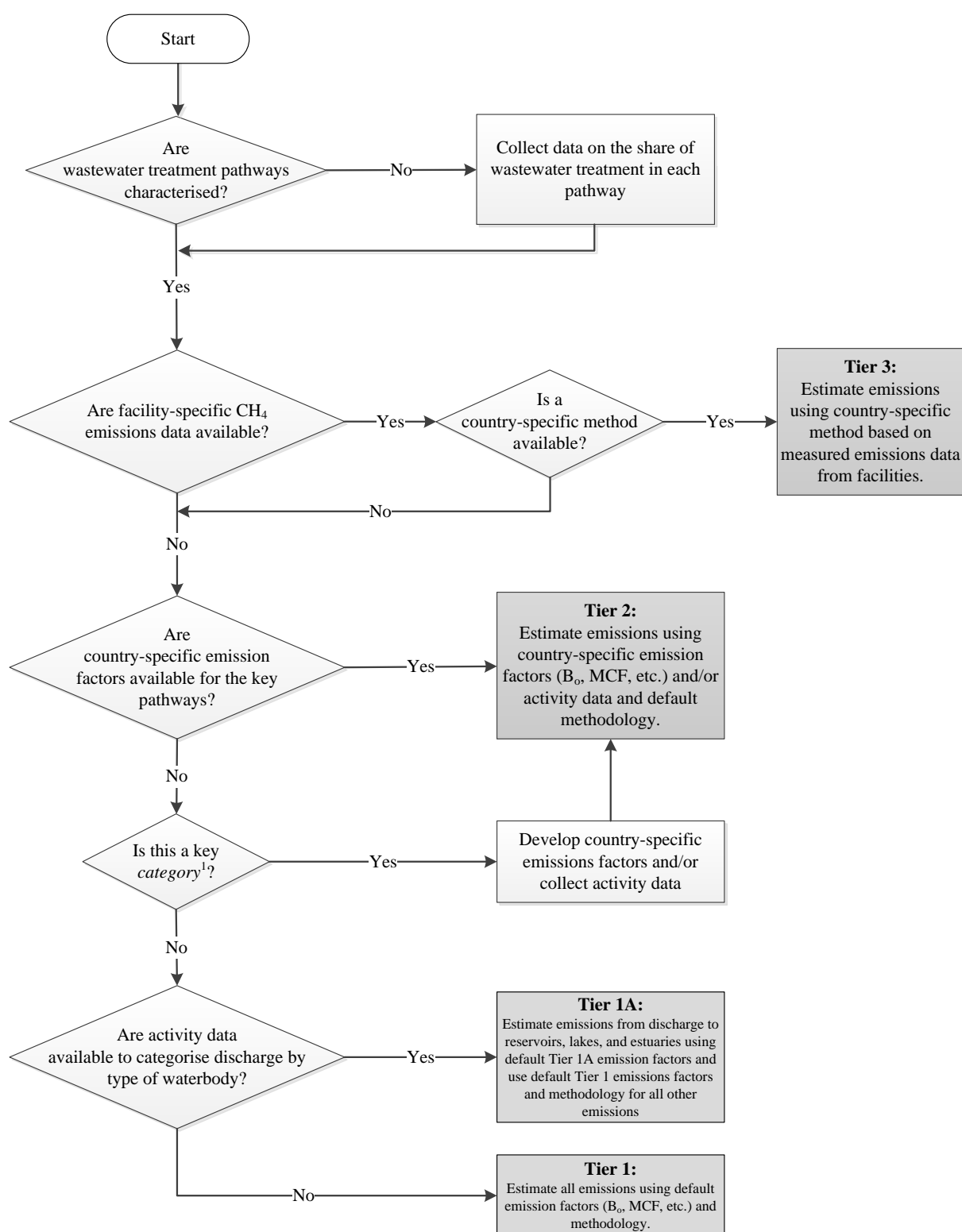
Emissions from flaring are not significant, as the majority of CO₂ emissions are of biogenic origin, and the CH₄ and N₂O emissions are very small so *good practice* in the Waste Sector does not require their estimation. However, if it is wished to do so these emissions should be reported under the Waste Sector. A discussion of emissions from flares and more detailed information are given in Volume 2, Energy, Chapter 4.2. Emission from flaring is not treated at Tier 1.

6.2.2 Domestic wastewater

6.2.2.1 CHOICE OF METHOD

An updated decision tree for domestic wastewater is included as updated Figure 6.2 and should be used to determine the tier approach that is applicable to the country.

This section is an update to the *2006 IPCC Guidelines*. In general, the overall steps for *good practice* in inventory preparation for CH₄ from domestic wastewater have been updated as follows:

368 **Figure 6.2 (Updated) Decision tree for CH₄ emissions from domestic wastewater**

¹ See Volume 1 Chapter 4, "Methodological Choice and Identification of Key Categories" (noting Section 4.1.2 on limited resources), for discussion of key categories and use of decision trees.

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Step 1: Use updated Equation 6.3 to estimate total organically degradable material in domestic wastewater (TOW). Use new Equation 6.3a to estimate TOW in wastewater prior to treatment for each wastewater treatment/discharge pathway or system, j , and each income group fraction i in inventory year.

Step 1A: Use new Equations 6.3b and 6.3c to estimate the amount of organic component removed in sludge, S , from aerobic treatment plants and septic systems for each wastewater treatment/discharge pathway or system, j , and each income group fraction, i , in the inventory year.

Step 1B: Use new Equation 6.3d to estimate total organics in treated wastewater effluent discharged ($TOW_{EFFtrat}$).

Step 2: Select the pathway and systems (See updated Figure 6.1) according to country activity data. Use Equation 6.2 or the updated Table 6.3 to obtain the emission factor for each domestic wastewater treatment/discharge pathway or system.

Step 3: Use updated Equation 6.1 to estimate emissions and adjust for possible sludge removal and/or CH_4 recovery of treatment/discharge pathway or system, j , in inventory year. Use new Equation 6.1a to sum the emissions across all treatment/discharge pathways or systems and to include emissions from the discharge of treated effluent.

Step 4: Use Equation 4.1 and emission factors in Table 4.1 to estimate methane emissions from anaerobic digestion of sludge.

To determine the use of each type of treatment or discharge system, it is *good practice* to refer to national statistics (e.g., from regulatory authorities). If these data are not available, wastewater associations or international organisations such as the World Health Organization (WHO) may have data on the system usage.

Otherwise, consultation with sanitation experts can help, and expert judgment can also be applied (see Chapter 2, Approaches to Data Collection, in Volume 1). Urbanisation statistics may provide a useful tool, e.g., city sizes and income distribution.

These updates are presented because, in some cases, the original Equation 6.1 caused confusion among inventory compilers when calculating CH_4 emissions across multiple wastewater treatment/discharge pathways or systems across multiple income groups. These updates allow for a more stepwise process in estimating emissions throughout the country.

Additional confusion arose in the estimation of emissions from aerobic systems with onsite anaerobic sludge digestion. In these cases, some compilers estimated zero CH_4 emissions from well operated wastewater treatment systems, and then subtracted CH_4 recovery associated with sludge digestion operations without first estimating the CH_4 emissions from onsite sludge treatment, resulting in negative emissions.

It is *good practice* for countries to treat the wastewater treatment system and onsite sludge treatment system as separate pathways. As an example, for an activated sludge treatment process, calculate the emissions directly associated with the aerobic treatment system as one pathway, and calculate the emissions and report any CH_4 recovery directly associated with the onsite sludge digestion system as a separate pathway using the emissions methodology provided in Table 4.1, Chapter 4 of Volume 5. Net emissions from both systems should be summed together. In no circumstances should a country report negative emissions.

It is important that CH_4 emissions from sludge that is managed offsite from the wastewater treatment facility using landfills, incineration, composting, biogas production, or used in agriculture are not included in the wastewater treatment and discharge category. The data should be consistent across the sectors, and categories, amount disposed at SWDS, applied to agricultural land, incinerated or used elsewhere should be equal to the amount organic component removed as sludge in updated Equation 6.1. Wastewater and sludge that is applied on agricultural land should be considered in Volume 4 for AFOLU Sector, Section 11.2, N_2O emissions from managed soils, in Chapter 11, N_2O Emissions from Managed Soils, and CO_2 Emissions from Lime and Urea Application.

Wastewater treatment system/pathway usage often differs for rural and urban residents. Also, in developing countries, there are likely to be differences between urban high-income and urban low-income residents. Hence, emissions are calculated by each income group fraction. It is *good practice* to treat the three categories: rural population, urban high income population, and urban low income population separately. It is suggested to use a spreadsheet, as shown in Table 6.5.

Equation 6.1 should be used for every treatment and discharge pathway or system that is appropriate for the country, including the discharge of treated or untreated wastewater. For the discharge of treated wastewater, the TOW should reflect the organics in the wastewater as discharged. Please see Section 6.2.2.3 for additional guidance on how to estimate S , organic component removed as sludge, if country-specific data are not available.

EQUATION 6.1 (UPDATED)
CH₄ EMISSIONS FROM DOMESTIC WASTEWATER FOR EACH TREATMENT/DISCHARGE PATHWAY
OR SYSTEM, *J*

$$CH_4 \text{ Emissions}_j = \left[(TOW_j - S_j) \bullet EF_j - R_j \right]$$

Where:

- $CH_4 \text{ Emissions}_j$ = CH_4 emissions from treatment/discharge pathway or system, *j*, in inventory year, kg CH_4 /yr
- TOW_j = organics in wastewater of treatment/discharge pathway or system, *j*, in inventory year, kg BOD/yr. See Equation 6.3a.
- S_j = organic component removed as sludge from treatment/discharge pathway or system, *j*, in inventory year, kg BOD/yr. See Equations 6.3b and 6.3c.
- j* = each treatment/discharge pathway or system
- EF_j = emission factor for treatment/discharge pathway or system, *j*, kg CH_4 /kg BOD. See Equation 6.2 or updated Table 6.3.
- R_j = amount of CH_4 recovered or flared from treatment/discharge pathway or system, *j*, in inventory year, kg CH_4 /yr. Default value is zero.

EQUATION 6.1A (NEW)
TOTAL CH₄ EMISSIONS FROM DOMESTIC WASTEWATER TREATMENT AND DISCHARGE

$$CH_4 \text{ Emissions} = [TOW_{EFF \text{ treat}} \bullet \% \text{ Tier 1} \bullet EF(\text{tier 1})] + [TOW_{EFF \text{ treat}} \bullet \% \text{ Tier 1A} \bullet EF(\text{tier 1A})] + \sum_j [CH_4 \text{ Emissions}_j]$$

Where:

- $CH_4 \text{ Emissions}$ = CH_4 emissions in inventory year, kg CH_4 /yr
- $TOW_{EFF \text{ treat}}$ = total organics in the treated wastewater effluent discharged to aquatic environments in inventory year, kg BOD/yr. See equation 6.3d.
- % Tier 1 = % wastewater discharged to locations other than reservoirs, lakes, and estuaries (default value is 100%)
- % Tier 1A = % wastewater discharged to reservoirs, lakes, and estuaries (default value is zero)
- EF = emission factor for discharge other than to reservoirs, lakes, and estuaries (Tier 1) or discharge to reservoirs, lakes, and estuaries (Tier 1a). See Equation 6.2 or updated Table 6.3.
- $CH_4 \text{ Emissions}_j$ = CH_4 emissions from treatment/discharge pathway or system, *j*, in inventory year, kg CH_4 /yr
- j* = each treatment/discharge pathway or system

6.2.2.2 CHOICE OF EMISSION FACTORS

This section represents an update to Section 6.2.2.2 of the 2006 IPCC Guidelines.

The emission factor for a wastewater treatment and discharge pathway and system is a function of the maximum CH_4 producing potential (B_o) and the methane correction factor (MCF) for the wastewater treatment and discharge system, as shown in Equation 6.2. The B_o is the maximum amount of CH_4 that can be produced from a given quantity of organics (as expressed in BOD or COD) in the wastewater and represents the complete conversion of organic C to biogas. The MCF indicates the extent to which the CH_4 producing capacity (B_o) is realised in each type of treatment and discharge pathway and system. Thus, it is an indication of the degree to which the system is anaerobic.

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$$\begin{array}{c}
 \text{EQUATION 6.2} \\
 \text{CH}_4 \text{ EMISSION FACTOR FOR} \\
 \text{EACH DOMESTIC WASTEWATER TREATMENT/DISCHARGE PATHWAY OR SYSTEM} \\
 EF_j = B_o \bullet MCF_j
 \end{array}$$

Where:

 EF_j = emission factor, kg CH₄/kg BOD

 j = each treatment/discharge pathway or system

 B_o = maximum CH₄ producing capacity, kg CH₄/kg BOD

 MCF_j = methane correction factor (fraction), See updated Table 6.3.

Table 6.2 includes default maximum CH₄ producing capacity (B_o) for domestic wastewater. It is *good practice* to use country-specific data if available for emission factors, which are made up of B_o and MCF values. If a country chooses to introduce country-specific data for B_o based on independent wastewater analyses, they must also update the MCF because the MCFs presented in Table 6.3 were developed using the default B_o values. For domestic wastewater, a COD-based value of B_o can be converted into a BOD-based value by multiplying with a factor of 2.4. New Annex 6A.2 provides further explanation of the basis for these default B_o values to allow countries to consider if these values are appropriate for the specific characteristics of their waste streams.

TABLE 6.2 DEFAULT MAXIMUM CH ₄ PRODUCING CAPACITY (B_o) FOR DOMESTIC WASTEWATER
0.6 kg CH ₄ /kg BOD
0.25 kg CH ₄ /kg COD
Based on expert judgment by lead authors and on Doorn <i>et al.</i> (1997)

Because the B_o and MCF values must be used together, updated Table 6.3 now also includes the resultant default CH₄ emission factors for each wastewater treatment and discharge pathway. In addition, the MCFs in Table 6.3 of the 2006 IPCC Guidelines have been updated to reflect revisions to the following specific wastewater treatment and discharge pathways and systems (see also new Annex 6A.3).

DISCHARGE FROM TREATED OR UNTREATED SYSTEMS

Updated Table 6.3 presents updated default MCFs associated with the discharge of wastewater to a water body and it is *good practice* to apply the Tier 1 MCF to discharges of both treated and untreated wastewater. The BOD of treated wastewater is typically 5–25 mg/L (Hammer & Hammer Jr. 2012; Tchobanoglous *et al.* 2014), which provides a substrate for the formation of CH₄ in a receiving water body. Furthermore, much of the dissolved CH₄ that is generated in sewers is released either in the treatment plant headworks or to the receiving water body in the case of untreated or primary treated discharges (Short *et al.* 2014; Short *et al.* 2017). Recent evidence points to the operation of both microbial and non-microbial methanogenic pathways in nature (e.g., Jugold *et al.* 2012) and strong relationships between the nutrient status of a receiving water body and the rate of generation of CH₄ have been observed (e.g., Smith *et al.* 2017). Despite this relationship, most rivers, estuaries and coastal waters are considerably supersaturated with CH₄, irrespective of their nutrient status (Patra *et al.* 1998; Grunwald *et al.* 2009; Ward *et al.* 2017), while open oceans are slightly supersaturated (Tilbrook & Karl 1995; Oudot *et al.* 2002; Castro-Morales *et al.* 2014). Supersaturated conditions occur when the rate of methanogenesis exceeds the rate at which the CH₄ is oxidised and/or transferred to the atmosphere and is important as it governs the driving force (mass transfer coefficient) and likely water-to-air CH₄ emission from a receiving environment. So, while the conditions of the receiving water body play a modulating role in relation to the rate of methanogenesis, the addition of organic matter from sewer discharges is generally expected to increase CH₄ emissions in freshwater and coastal environments. Recent reviews of measurements of CH₄ generation indicate that a significant proportion of the CH₄ emitted from freshwater systems has its origins in carbon deposited on sediments (Deemer *et al.* 2016) which is one reason why methanogenesis is more intense in lakes and reservoirs than fast-flowing rivers.

Using the same stoichiometric relationship as was used to calculate the default B_o value (see new Annex 6A.2), a ratio of 0.938 kg C per kg COD is obtained. Deemer *et al.* (2016) performed an extensive review of measurements of CH₄ and CO₂ originating in aquatic systems, and independent measurements of both CH₄ and CO₂ emission fluxes around the world. Using these data in combination with information on the partitioning of global carbon flows in freshwater systems Tranvik *et al.* (2009) and a default B_o of 0.25 kg CH₄/kg COD, an MCF of 0.035 is calculated for rivers. This is also the default Tier 1 MCF for CH₄ generation for discharges to freshwater and

coastal environments. However, where it is possible to separately identify discharges to slow-flowing systems such as lakes and reservoirs, a Tier 1a MCF of 0.19 is presented (see updated Table 6.3 and new Annex 6A.4).

Concerning decentralised treatment systems for domestic wastewater, CH₄ and N₂O emissions from effluent infiltration into soil must be considered. Default CH₄ and N₂O EFs related to discharge in non-aquatic environments are proposed.

CENTRALISED, AEROBIC TREATMENT PLANTS

The MCF for centralised aerobic treatment plants has been updated in Table 6.3 to reflect the potential for generation of CH₄ from these systems. In addition, there are no longer separate MCFs for “well managed” and “not well managed” systems and it is *good practice* to estimate CH₄ from all centralised, aerobic treatment plants. If country-specific data are available to differentiate whether systems are overloaded or not well managed, these situations should be reflected in the calculation of TOW (for inflow overload) or S_{mass} (for systems that are not well managed and therefore not achieving the expected removal of sludge).

IMPACT OF TEMPERATURE ON DECENTRALISED TREATMENT SYSTEMS OF DOMESTIC WASTEWATER

Temperature affects wastewater treatment processes, in particular decentralised systems where no external supplemental heat is provided (uncontrolled temperature) and anaerobic digestion for which the optimal temperature is 30–38°C. At lower temperatures, the rate of anaerobic digestion decreases and CH₄ production becomes unlikely below 12°C. Inside septic tanks, the temperature is uncontrolled and is related to atmospheric temperature as well as volumes of household hot and cold water used and discharged. There may also be a gradient of temperature inside the septic tank, with warmer conditions at the bottom (sludge layer) and colder at the top (Leverenz *et al.* 2010). Therefore, in countries having seasonal temperature variability, when the temperature in septic tanks cools, the rate of digestion slows, the SRT increases, sludge accumulates, and CH₄ emissions decrease. When the liquid temperature warms, the rate of digestion increases, sludge accumulated during the cold season decomposes, gas solubility in the liquid decreases and CH₄ emissions increase. This situation can produce a ‘spring boil’ phenomenon, wherein warmer weather conditions give rise to increased anaerobic microbial activity, increased gas production, and decreased solids removal efficiency due to the resuspension of settled and incoming solids. Accordingly, there is a seasonal variability of CH₄ emissions (Leverenz *et al.* 2010); however, at this time, insufficient data exist to establish a temperature-dependent emission factor associated with these systems. Countries that experience significant seasonal temperature variations may wish to consider the development of a country-specific emission factor.

TABLE 6.3 (UPDATED)				
DEFAULT MCF VALUES AND RESULTANT EF ² FOR DOMESTIC WASTEWATER ¹				
Type of treatment and discharge pathway or system	Comments	MCF ¹ (Range)	EF ² (kg CH ₄ /kg BOD)	EF ² (kg CH ₄ /kg COD)
Discharge from treated or untreated system				
Discharge other than to reservoirs, lakes, and estuaries (Tier 1)	Most freshwater systems including rivers are supersaturated in CH ₄ . Nutrient oversupply will increase CH ₄ emissions.	0.035 ³ (0.004 – 0.06)	0.021	0.009
Discharge to reservoirs, lakes, and estuaries (Tier 1a)	Environments where carbon accumulates in sediments have higher potential for methane generation.	0.19 ³ (0.08 – 0.27)	0.114	0.048
Discharge to soil	Dry climate; negligible emissions	0	0	0
Discharge to soil	Wet climate	0.7 (0.7 – 1.0)	0.42	0.175
Untreated system				
Stagnant sewer	Open and warm	0.5 (0.4 – 0.8)	0.3	0.125
Flowing sewer (open or closed)	Fast moving, clean. (Insignificant amounts of CH ₄ from pump stations, etc.)	0	0	0

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TABLE 6.3 (UPDATED) (CONTINUED)				
DEFAULT MCF VALUES AND RESULTANT EF ² FOR DOMESTIC WASTEWATER ¹				
Type of treatment and discharge pathway or system	Comments	MCF ¹ (Range)	EF ² (kg CH ₄ /kg BOD)	EF ² (kg CH ₄ /kg COD)
Treated wastewater treatment system				
Centralised, aerobic treatment plant	Some CH ₄ can be emitted from settling basins and other anaerobic pockets. May also emit CH ₄ generated in upstream sewer networks during turbulent and/or aerobic treatment processes. For treatment plants that are receiving wastewater beyond the design capacity, inventory compilers should judge the amount of organic material removed in sludge accordingly.	0.03 ⁴ (0.003 – 0.09)	0.018	0.0075
Anaerobic reactor (e.g., upflow anaerobic sludge blanket digestion (UASB))	CH ₄ recovery is not considered here.	0.8 (0.8 – 1.0)	0.48	0.2
Anaerobic shallow lagoon and facultative lagoons	Depth less than 2 metres, use expert judgment.	0.2 (0 – 0.3)	0.12	0.05
Anaerobic deep lagoon	Depth more than 2 metres	0.8 (0.8 – 1.0)	0.48	0.2
Constructed wetlands	See 2013 Supplement to the 2006 IPCC Guidelines for National Greenhouse Gas Inventories: Wetlands (IPCC 2014)			
Septic tank	Septic tanks emit CH ₄	0.5 ⁵ (0.4 – 0.72)	0.3	0.125
Septic tank + land dispersal field	Septic tanks emit CH ₄ ; negligible emissions come from land dispersal field	0.5 ⁵ (0.4 – 0.72)	0.3	0.125
Latrine	Dry climate, ground water table lower than latrine, small family (3–5 persons)	0.1 (0.05 – 0.15)	0.06	0.025
Latrine	Dry climate, ground water table lower than latrine, communal (many users)	0.5 (0.4 – 0.6)	0.3	0.125
Latrine	Wet climate/flush water use, ground water table higher than latrine	0.7 (0.7 – 1.0)	0.42	0.175
Latrine	Regular sediment removal for fertiliser	0.1	0.06	0.025
Onsite sludge treatment system				
Anaerobic digester for sludge	See Chapter 4 for emissions methodology	See Chapter 4, Table 4.1		
Composting	See Chapter 4 for emissions methodology	See Chapter 4, Table 4.1		
Incineration and open burning	See Chapter 5 for emissions methodology	See Chapter 5		
Sources:				
¹ Based on expert judgment by Lead Authors of this section.				
² Emission factors calculated using default B _o and default MCF.				
³ See Annex 6A.4.				
⁴ See Annex 6A.3 (Czepiel <i>et al.</i> 1993; Kozak <i>et al.</i> 2009; Bellucci <i>et al.</i> 2010; Wang <i>et al.</i> 2011; Daelman <i>et al.</i> 2013; Kyung <i>et al.</i> 2015; Delre <i>et al.</i> 2017).				
⁵ Leverenz <i>et al.</i> 2010; Diaz-Valbuena <i>et al.</i> 2011; Truhlar <i>et al.</i> 2016.				

6.2.2.3 CHOICE OF ACTIVITY DATA

The activity data for this source category is the total amount of organically degradable material in the wastewater (TOW). This parameter is a function of human population and BOD generation per person. It is expressed in terms of biochemical oxygen demand (kg BOD/year). The equation for TOW is:

$$\text{EQUATION 6.3 (UPDATED)}$$

$$\text{TOTAL ORGANICALLY DEGRADABLE MATERIAL IN DOMESTIC WASTEWATER}$$

$$TOW = P \bullet BOD \bullet 0.001 \bullet 365$$

Where:

TOW = total organics in wastewater in inventory year, kg BOD/yr

P = country population in inventory year, (person)

BOD = country-specific per capita BOD in inventory year, g/person/day, See Table 6.4.

0.001 = conversion from grams BOD to kg BOD

Total population statistics should be readily available from national statistics agencies or international agencies (e.g., United Nations Statistics, see <https://population.un.org/wpp/>). Table 6.4 includes BOD default values for selected countries. It is *good practice* to select a BOD default value from a nearby comparable country when country-specific data are not available. The degree of urbanization for a country can be retrieved from various sources, (e.g., Global Environment Outlook, United Nations Environment Programme and World Development Indicators, World Health Organization). The urban high-income and urban-low income fractions can be determined by expert judgment when statistical or other comparable information is not available.

TABLE 6.4 ESTIMATED BOD ₅ VALUES IN DOMESTIC WASTEWATER FOR SELECTED REGIONS AND COUNTRIES			
Country/Region	BOD ₅ (g/person/day)	Range	Reference
Africa	37	35 – 45	1
Egypt	34	27 – 41	1
Asia, Middle East, Latin America	40	35 – 45	1
India	34	27 – 41	1
West Bank and Gaza Strip (Palestine)	50	32 – 68	1
Japan	42	40 – 45	1
Brazil	50	45 – 55	2
Canada, Europe, Russia, Oceania	60	50 – 70	1
Denmark	62	55 – 68	1
Germany	62	55 – 68	1
Greece	57	55 – 60	1
Italy	60	49 – 60	3
Sweden	75	68 – 82	1
Turkey	38	27 – 50	1
United States	85	50 – 120	4
Note: These values are based on an assessment of the literature. Please use national values, if available. Reference: 1. Doorn and Liles (1999). 2. Feachem <i>et al.</i> (1983). 3. Masotti (1996). 4. Metcalf and Eddy (2003).			

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This section is updated to include a new equation for the calculation of total organics in wastewater (TOW_j) by treatment/discharge pathway or system (see new Equation 6.3a).

EQUATION 6.3A (NEW)
TOTAL ORGANICS IN DOMESTIC WASTEWATER BY TREATMENT/DISCHARGE PATHWAY OR SYSTEM

$$TOW_j = \sum_i [TOW \bullet U_i \bullet T_{ij} \bullet I_j]$$

Where:

TOW_j = total organics in wastewater in inventory year, kg BOD/yr, for income group i and treatment/discharge pathway or system, j .

TOW = total organics in wastewater in inventory year, kg BOD/yr. See updated Equation 6.3 for TOW in wastewater prior to treatment and new Equation 6.3d for TOW in wastewater effluent.

U_i = fraction of population in income group i in inventory year. See Table 6.5.

T_{ij} = degree of utilisation of treatment/discharge pathway or system, j , for each income group fraction

I = correction factor for additional industrial BOD discharged into sewers (for collected the default is 1.25, for uncollected the default is 1.00)

The factor I values in new Equation 6.3a are from the 2006 IPCC Guidelines. It expresses the BOD from industries and establishments (e.g., restaurants, butchers or grocery stores) that is co-discharged with domestic wastewater. In some countries, information from industrial discharge permits may be available to improve I . Otherwise, expert judgment is recommended.

Table 6.5 includes default values of U_i and T_{ij} for selected countries.

Example

Table 6.6 includes an example. Categories with negligible contributions are not shown. Note that the table can easily be expanded with a column for MCF for each category. The degree of urbanization for this country is 65 percent.

TABLE 6.6 EXAMPLE OF THE APPLICATION OF DEFAULT VALUES FOR DEGREES OF TREATMENT UTILIZATION (T) BY INCOME GROUPS			
Treatment or discharge system or pathway		T (%)	Notes
Urban high-income	To sea	10	No CH ₄
	To aerobic plant	20	Add industrial component
	To septic systems	10	Uncollected
Urban low-income	To sea	10	Collected
	To pit latrines	15	Uncollected
Rural	To rivers, lakes, sea	15	Uncollected
	To pit latrines	15	
	To septic tanks	5	
Total		100%	Must add up to 100%
Reference: Doorn and Liles (1999)			

This section is updated to also include new equation for the calculation of organic component removed as sludge from aerobic treatment plants and from septic systems (see Equations 6.3b and 6.3c, respectively). Note that the estimate of TOW_{ij} in Equation 6.3a is specific to the wastewater treatment system or pathway. Inventory compilers should consider that sludge recovered from septic tanks may be transferred to centralised WWTPs. In these cases,

599 it is *good practice* to include this additional organic load when estimating TOW in influent to the centralised
600 WWTP.

601 The organic component removed from wastewater as sludge, S , in Equations 6.1 and 6.4, is not explained in detail
602 in the 2006 IPCC Guidelines. This section provides an update through the introduction of new Equations 6.3b and
603 6.3c, which provide default calculations of S for aerobic treatment plants and septic systems, respectively. The
604 default value of S for all other systems is zero. For aerobic treatment systems, some inventory compilers incorrectly
605 defined the variable “ S ” in Equation 6.1 as the mass of sludge removed rather than the organic component removed
606 from wastewater as sludge. It is important to note that the organic component removed from wastewater as sludge
607 is not equivalent to the mass (tonnes) of sludge produced from wastewater treatment. Instead, the organic
608 component removed as sludge is a function of sludge produced from wastewater treatment (S_{MASS}) and a sludge
609 factor (K_{REM}) which indicates how much organic matter is removed from the treatment process in sludge per ton
610 of sludge produced.

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	Urbanization(U) ¹			Degree of utilisation of treatment or discharge pathway or method for each income group (T_{ij}) ³														
	Fraction of Population			U=rural					U= urban high income					U=urban low income				
Country	Rural	urban-high²	urban-low²	Septic Tank	Latrine	Other	Sewer⁴	None	Septic Tank	Latrine	Other	Sewer⁴	None	Septic Tank	Latrine	Other	Sewer⁴	None
Africa																		
Nigeria	0.52	0.10	0.38	0.02	0.28	0.04	0.10	0.56	0.32	0.31	0.00	0.37	0.00	0.17	0.24	0.05	0.34	0.20
Egypt	0.57	0.09	0.34	0.02	0.28	0.04	0.10	0.56	0.15	0.05	0.10	0.70	0.00	0.17	0.24	0.05	0.34	0.20
Kenya	0.62	0.08	0.30	0.02	0.28	0.04	0.10	0.56	0.32	0.31	0.00	0.37	0.00	0.17	0.24	0.05	0.34	0.20
South Africa	0.39	0.12	0.49	0.10	0.28	0.04	0.10	0.48	0.15	0.15	0.00	0.70	0.00	0.17	0.24	0.05	0.34	0.20
Asia																		
China	0.59	0.12	0.29	0.00	0.47	0.50	0.00	0.3	0.18	0.08	0.07	0.67	0.00	0.14	0.10	0.03	0.68	0.05
India	0.71	0.06	0.23	0.00	0.47	0.10	0.10	0.33	0.18	0.08	0.07	0.67	0.00	0.14	0.10	0.03	0.53	0.20
Indonesia	0.54	0.12	0.34	0.00	0.47	0.00	0.10	0.43	0.18	0.08	0.00	0.74	0.00	0.14	0.10	0.03	0.53	0.20
Pakistan	0.65	0.07	0.28	0.00	0.47	0.00	0.10	0.43	0.18	0.08	0.00	0.74	0.00	0.14	0.10	0.03	0.53	0.20
Bangladesh	0.72	0.06	0.22	0.00	0.47	0.00	0.10	0.43	0.18	0.08	0.00	0.74	0.00	0.14	0.10	0.03	0.53	0.20
Japan	0.20	0.80	0.00	0.20	0.00	0.50	0.30	0.00	0.00	0.00	0.10	0.90	0.00	0.10	0	0	0.90	0
Europe																		
Russia	0.27	0.73	0.00	0.30	0.10	0.00	0.60	0.00	0.10	0.00	0.00	0.90	0.00	NA	NA	NA	NA	NA
Germany ⁵	0.06	0.94	0.00	0.20	0.00	0.00	0.80	0.00	0.05	0.00	0.00	0.95	0.00	NA	NA	NA	NA	NA
United Kingdom	0.10	0.90	0.00	0.11	0.00	0.00	0.89	0.00	0.00	0.00	0.00	1.00	0.00	NA	NA	NA	NA	NA
France	0.24	0.76	0.00	0.37	0.00	0.00	0.63	0.00	0.00	0.00	0.00	1.00	0.00	NA	NA	NA	NA	NA
Italy	0.32	0.68	0.00	0.42	0.00	0.00	0.58	0.00	0.04	0.00	0.00	0.96	0.00	NA	NA	NA	NA	NA
North America																		
United States	0.22	0.78	0.00	0.90	0.02	0.00	0.08	0.00	0.05	0.00	0.00	0.95	0.00	NA	NA	NA	NA	NA
Canada	0.20	0.80	0.00	0.90	0.02	0.00	0.08	0.00	0.05	0.00	0.00	0.95	0.00	NA	NA	NA	NA	NA
Latin America and Caribbean																		
Brazil	0.16	0.25	0.59	0.00	0.45	0.00	0.10	0.45	0.00	0.20	0.00	0.80	0.00	0.00	0.40	0.00	0.40	0.20
Mexico	0.25	0.19	0.56	0.00	0.45	0.00	0.10	0.45	0.00	0.20	0.00	0.80	0.00	0.00	0.40	0.00	0.40	0.20
Oceania																		
Australia and New Zealand	0.08	0.92	0.00	0.90	0.02	0.00	0.08	0.00	0.05	0.00	0.00	0.95	0.00	NA	NA	NA	NA	NA
Notes:																		
1. Urbanization projections for 2005 (United Nations, 2002).																		
2. Suggested urban-high income and urban low income division. Countries are encouraged to use their own data or best judgment.																		
3. T _{ij} values based on expert judgment, (Doorn and Liles, 1999).																		
4. Sewers may be open or closed, which will govern the choice of MCF, see Table 3.3																		
5. Destatis, 2001.																		
Note: These values are from the literature or based on expert judgment. Please use national values, if available.																		

EQUATION 6.3B (NEW)
ORGANIC COMPONENT REMOVED AS SLUDGE FROM AEROBIC TREATMENT PLANTS

$$S_{aerobic} = (S_{mass} \bullet K_{rem} \bullet 1000)$$

Where:

- $S_{aerobic}$ = organic component removed as sludge in aerobic treatment plants, kg BOD/yr
 S_{mass} = amount of raw sludge removed from waste water treatment as dry mass, tonnes/year
 K_{rem} = sludge factor, kg BOD/kg sludge. See Table 6.6a.
1000 = conversion factor for tonnes to kilograms

New Table 6.6a provides factors associated with the amount of organics removed in sludge. Aerobic WWTPs with primary treatment only remove about 0.5 kg of influent BOD per kg of primary sludge by sedimentation. The aerobic stage of treatment removes 1.5 kg BOD per kg of secondary sludge. The sludge produced in aerobic WWTPs with primary sedimentation contains approximately 70% of primary sludge and 30% of secondary sludge by weight. Aerobic WWTPs without primary treatment remove about 1.16 kg BOD per kg of sludge, depending on process type. Countries where data are available should estimate a country-specific K_{REM} as a weighted average of sludge generated by individual processes.

TABLE 6.6A (NEW) REMOVAL OF ORGANIC COMPONENT FROM WASTEWATER AS SLUDGE (K_{REM}) ACCORDING TO TREATMENT TYPE ¹		
Treatment Type	Default	Range
	(kg BOD/kg dry mass sludge)	
Mechanical treatment plants (primary sedimentation sludge)	0.5	0.4 – 0.6
Aerobic treatment plants with primary treatment (mixed primary and secondary sludge, untreated or treated aerobically)	0.8	0.65 – 0.95
Aerobic treatment plants with primary treatment and anaerobic sludge digestion (mixed primary and secondary sludge, treated anaerobically)	1.0	0.8 – 1.2
Aerobic wastewater treatment plants without separate primary treatment	1.16	1.0 – 1.5
Sources: ¹ Based on expert judgment by Lead Authors of this section using the following references: Pescod (1992); Davies (2005); Foladori <i>et al.</i> (2010); WEF (2010); Wisconsin Department of Natural Resources (2010); Serón <i>et al.</i> (2011).		

Concerning septic systems, emissions depend on the fraction of the population managing their septic tank in accordance with the sludge removal instructions of their septic tank/system (F in new Equation 6.3c). The default value for F is 0.5 and corresponds to the situation where 50% of the population managing their septic system are complying with the sludge removal instructions. This default value is for countries where there is no effective regulation or administrative requirements for sludge removal in septic systems. In countries with such regulations or requirements, some evidence of maintenance controls should be provided (for instance the existence of a local public service responsible for onsite sanitation). It is *good practice* to assess the F value using available data on sludge removal practices among the population using septic systems.

EQUATION 6.3C (NEW)
ORGANIC COMPONENT REMOVED AS SLUDGE FROM SEPTIC SYSTEMS

$$S_{septic} = TOW_{septic} \bullet F \bullet 0.5$$

Where:

- S_{septic} = organic component removed as sludge in septic systems, kg BOD/yr
 TOW_{septic} = total organics in wastewater in septic systems inventory year, kg BOD/yr
 F = fraction of the population managing their septic tank in compliance with the sludge removal instruction of their septic system

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0.5 = fraction of organics in wastewater removed in sludge when septic tank is managed in accordance with sludge removal instructions

For discharges of wastewater from treated or untreated systems, inventory compilers should estimate the amount of TOW in effluent ($TOW_{EFFLUENT}$) as shown in new Equation 6.3d.

EQUATION 6.3D (NEW)
TOTAL ORGANICS IN TREATED DOMESTIC WASTEWATER EFFLUENT

$$TOW_{EFFtreat} = \sum_k \left[TOW \cdot FWT_k \cdot (1 - TOW_{REM,k}) \right]$$

Where:

$TOW_{EFFtreat}$ = total organics in the treated wastewater effluent discharged to aquatic environments in inventory year, kg BOD/yr

TOW = total organically degradable material in domestic wastewater in inventory year, kg BOD/yr. See Equation 6.3 (Updated).

FWT_k = fraction of wastewater treated exclusively by each wastewater treatment type k in inventory year. See Table 6.5.

k = each wastewater treatment type used in inventory year

TOW_{REM} = fraction of total wastewater organics removed during wastewater treatment per treatment type k . See new Table 6.6b. Pathways for organics removal include loss to sludge and biological decomposition.

TABLE 6.6B (NEW)
WASTEWATER TREATMENT ORGANICS REMOVAL FRACTIONS (TOW_{REM}) ACCORDING TO TREATMENT TYPE

Treatment Type	Default	Range
Primary (mechanical)	0.40	0.25 – 0.50
Secondary (biological)	0.85	0.80 – 0.90
Tertiary (advanced biological)	0.90	0.80 – 0.95
Septic tank/septic system	0.625	0.50 – 0.60
Latrines – Dry climate, groundwater table lower than latrine, small family (3–5 persons)	0.1	0.05 – 0.15
Latrines – Dry climate, groundwater table lower than latrine, communal (many users)	0.5	0.4 – 0.6
Latrines – Wet climate/flush water use, groundwater table higher than latrine	0.7	0.7 – 1.0
Latrines – Regular sediment removal for fertiliser	0.1	0.05 – 0.15
Sources: ¹ Based on expert judgment by Lead Authors of this section using the following references: Pescod (1992); WEF (2009); Schaider <i>et al.</i> (2017).		

6.2.2.4 TIME SERIES CONSISTENCY

No refinement

6.2.2.5 UNCERTAINTIES

Chapter 3, Uncertainties, in Volume 1 provides advice on quantifying uncertainties in practice. It includes guidance on eliciting and using expert judgments which in combination with empirical data can provide overall uncertainty estimates. This section provides an update to Table 6.7 to provide default uncertainty ranges for new or updated emission factor and activity data of domestic wastewater. The following parameters are believed to be very uncertain:

- The degrees to which wastewater in developing countries is treated in latrines, septic tanks, or removed by sewer, for urban high, urban low income groups and rural population (T_{ij}).
- The fraction of sewers that are 'open', as well as the degree to which open sewers in developing countries are anaerobic and will emit CH_4 . This will depend on retention time and temperature, and on other factors including the presence of a facultative layer and possibly components that are toxic to anaerobic bacteria (e.g., certain industrial wastewater discharges).
- The amount of industrial TOW that is discharged into open or closed domestic sewers for each country is very difficult to quantify.

TABLE 6.7 (UPDATED) DEFAULT UNCERTAINTY RANGES FOR DOMESTIC WASTEWATER	
Parameter	Uncertainty Range
Emission Factor	
Maximum CH_4 producing capacity (B_o)	$\pm 30\%$
Fraction treated anaerobically (MCF)	The MCF is technology dependent. See Table 6.3. Thus, the uncertainty range is also technology dependent. The uncertainty range should be determined by expert judgment, bearing in mind that MCF is a fraction and must be between 0 and 1. Suggested ranges are provided below. Untreated systems, $\pm 50\%$ Lagoons $\pm 30\%$ Centralised plant, digester, reactor, $\pm 10\%$
Activity Data	
Human population (P)	$\pm 5\%$
Biochemical oxygen demand (BOD)	$\pm 30\%$
Fraction of population income group (U)	Good data on urbanization are available, however, the distinction between urban high income and urban low income may have to be based on expert judgment. $\pm 15\%$
Degree of utilisation of treatment/discharge pathway or system for each income group (T_{ij})	Can be as low as $\pm 3\%$ for countries that have good records and only one or two systems. Can be $\pm 50\%$ for an individual method/pathway. Verify that total $T_{ij} = 100\%$
Correction factor for additional industrial BOD discharged into sewers (I)	For uncollected, the uncertainty is zero %. For collected the uncertainty is $\pm 20\%$
Amount of sludge removed from waste water treatment (S_{mass})	$\pm 30\%$
Sludge factor (K_{rem})	$\pm 25\%$
Fraction of the population managing their septic system complying with the sludge removal instruction (F)	Can be as low as $\pm 3\%$ for countries that have good records on implementation. Can be $\pm 50\%$ if based on expert judgment.
Amount of CH_4 recovered or flared (R)	For systems with measured data, the uncertainty is equal to the uncertainty of the measurement system.
Source: Based on expert judgment by Lead Authors of this section.	

6.2.2.6 QA/QC, COMPLETENESS, REPORTING AND DOCUMENTATION

No refinement

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6.2.3 Industrial wastewater

Industrial wastewater may be treated on site or released into domestic sewer systems. If it is released into the domestic sewer system, the emissions are to be included with the domestic wastewater emissions. This section deals with estimating CH₄ emissions from on-site industrial wastewater treatment. Only industrial wastewater with significant carbon loading that is treated under intended or unintended anaerobic conditions will produce CH₄. Organics in industrial wastewater are often expressed in terms of COD, which is used here.

6.2.3.1 CHOICE OF METHOD

Assessment of CH₄ production potential from industrial wastewater streams is based on the concentration of degradable organic matter in the wastewater, the volume of wastewater, and the propensity of the industrial sector to treat their wastewater in anaerobic systems. Using these criteria, major industrial wastewater sources with high CH₄ gas production potential can be identified as follows:

- pulp and paper manufacture,
- meat and poultry processing (slaughterhouses),
- alcohol, beer, starch production,
- organic chemicals production,
- other food and drink processing (dairy products, vegetable oil, fruits and vegetables, canneries, juice making, etc.).

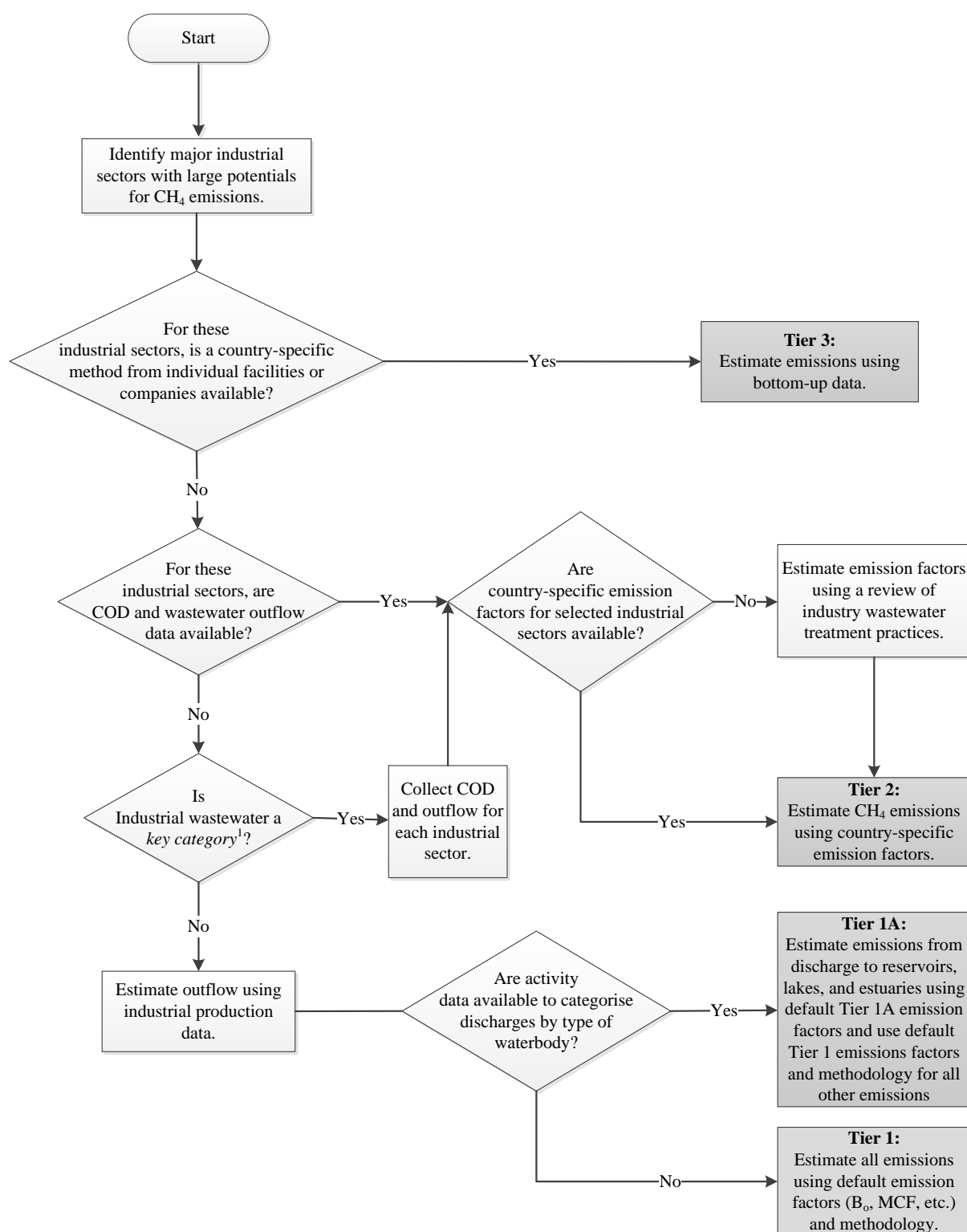
Both the pulp and paper industry and the meat and poultry processing industries produce large volumes of wastewater that contain high levels of degradable organics. The meat and poultry processing facilities typically employ anaerobic lagoons to treat their wastewater, while the paper and pulp industry also use lagoons and anaerobic reactors. The non-animal food and beverage industries produce considerable amounts of wastewater with significant organic carbon levels and are also known to use anaerobic processes such as lagoons and anaerobic reactors. Anaerobic reactors treating industrial effluents with biogas facilities are usually linked with recovery of the generated CH₄ for energy. Emissions from the combustion process for energy should be reported in the Energy Sector.

The method for estimating emissions from industrial wastewater is similar to the one used for domestic wastewater. An updated decision tree for industrial wastewater is included as updated Figure 6.3 and should be used to determine the tier approach that is applicable to the country.

The development of emission factors and activity data is more complex because there are many types of wastewater, and many different industries to track. The most accurate estimates of emissions for this source category would be based on measured data from point sources. Due to the high costs of measurements and the potentially large number of point sources, collecting comprehensive measurement data is very difficult. It is suggested that inventory compilers use a top-down approach that includes the following general steps:

- Step 1:** Use Equation 6.6 to estimate total organically degradable carbon in wastewater (TOW) for industrial sector *i*
- Step 2:** Select the pathway and systems (updated Figure 6.1) according to country activity data. Use Equation 6.5 to obtain emission factor. For each industrial sector estimate the emission factor using maximum methane producing capacity and the average industry-specific methane correction factor.
- Step 3:** Use Equation 6.4 to estimate emissions, adjust for possible sludge removal and or CH₄ recovery and sum the results.

725 **Figure 6.3 (Updated) Decision tree for CH₄ emissions from industrial wastewater treatment**



¹ See Volume 1 Chapter 4, "Methodological Choice and Identification of Key Categories" (noting Section 4.1.2 on limited resources), for discussion of key categories and use of decision trees.

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The general equation to estimate CH₄ emissions from industrial wastewater is as follows:

$$\text{TOTAL CH}_4 \text{ EMISSIONS FROM INDUSTRIAL WASTEWATER}$$

$$CH_4 \text{ Emissions} = \sum_i [(TOW_i - S_i) EF_i - R_i]$$

Where:

CH₄ Emissions = CH₄ emissions in inventory year, kg CH₄/yr

TOW_{*i*} = total organically degradable material in wastewater from industry *i* in inventory year, kg COD/yr

i = industrial sector

S_{*i*} = organic component removed as sludge in inventory year, kg COD/yr

EF_{*i*} = emission factor for industry *i*, kg CH₄/kg COD for treatment/discharge pathway or system(s) used in inventory year

If more than one treatment practice is used in an industry this factor would need to be a weighted average.

R_{*i*} = amount of CH₄ recovered or flared in inventory year, kg CH₄/yr

The amount of CH₄ which is recovered or flared is expressed as R in Equation 6.4.

6.2.3.2 CHOICE OF EMISSION FACTORS

This section represents an update to Section 6.2.3.2 of the 2006 IPCC Guidelines.

As stated in the 2006 IPCC Guidelines, there are significant differences in the CH₄ emitting potential of different types of industrial wastewater dependent on the type and form of constituents present in the wastewater. To the extent possible, data should be collected to determine the maximum CH₄ producing capacity (B_o) in each industry.

As mentioned before, the MCF indicates the extent to which the CH₄ producing potential (B_o) is realised in each type of treatment method. Thus, it is an indication of the degree to which the system is anaerobic. See Equation 6.5.

$$\text{CH}_4 \text{ EMISSION FACTOR FOR INDUSTRIAL WASTEWATER}$$

$$EF_j = B_o \bullet MCF_j$$

Where:

EF_{*j*} = emission factor for each treatment/discharge pathway or system, kg CH₄/kg COD. See Table 6.8.

j = each treatment/discharge pathway or system

B_o = maximum CH₄ producing capacity, kg CH₄/kg COD

MCF_{*j*} = methane correction factor (fraction). See Table 6.8.

Good practice is to use country- and industry-sector specific data that may be available from government authorities, industrial organisations, or industrial experts. If country-specific data are available to determine the B_o for a particular industry, industry-specific MCFs must be developed for the calculation of CH₄ emissions. However, if country-specific data are not available, it is *good practice* to use the IPCC default factors listed in Table 6.2.

In determining the Methane correction factor (MCF), which is the fraction of waste treated anaerobically, expert judgment is recommended. A peer-reviewed survey of industry wastewater treatment practices is one useful technique for estimating these data. Surveys should be conducted frequently enough to account for major trends in industry practices (i.e., every 3-5 years). Chapter 2, Approaches to Data Collection, in Volume 1, describes how to elicit expert judgment for uncertainty ranges. Similar expert elicitation protocols can be used to obtain the necessary information for other types of data if published data and statistics are not available. Table 6.8 includes default MCF values, which are based on expert judgment.

In addition, the MCFs in Table 6.8 have been updated to reflect revisions to certain wastewater treatment and discharge pathways and systems, as described in Section 6.2.2.2.

TABLE 6.8 (UPDATED) DEFAULT MCF VALUES AND RESULTANT EF ² FOR INDUSTRIAL WASTEWATER				
Type of treatment and discharge pathway or system	Comments	MCF ¹ (Range)	EF ² (kg CH ₄ /kg BOD)	EF ² (kg CH ₄ /kg COD)
Discharge from treated or untreated system				
Discharge other than to reservoirs, lakes, and estuaries (Tier 1)	Most freshwater systems including rivers are supersaturated in CH ₄ . Nutrient oversupply will increase CH ₄ emissions.	0.035 ³ (0.004 – 0.06)	0.021	0.009
Discharge to reservoirs, lakes, and estuaries (Tier 1a)	Environments where carbon accumulates in sediments have higher potential for methane generation.	0.19 ³ (0.08 – 0.27)	0.114	0.048
Treated wastewater treatment system				
Centralised, aerobic treatment plant	Some CH ₄ can be emitted from settling basins and other anaerobic pockets. For treatment plants that are receiving wastewater beyond the design capacity, inventory compilers should judge the amount of organic material removed in sludge accordingly.	0 (0 – 0.1)	0	0
Anaerobic reactor (e.g., upflow anaerobic sludge blanket digestion (UASB))	CH ₄ recovery is not considered here.	0.8 (0.8 – 1.0)	0.48	0.2
Anaerobic shallow lagoon and facultative lagoons	Depth less than 2 metres, use expert judgment.	0.2 (0 – 0.3)	0.12	0.05
Anaerobic deep lagoon	Depth more than 2 metres.	0.8 (0.8 – 1.0)	0.48	0.2
Constructed wetlands	See 2013 Supplement to the 2006 IPCC Guidelines for National Greenhouse Gas Inventories: Wetlands (IPCC 2014)			
Onsite sludge treatment system				
Anaerobic digester for sludge	See Chapter 4 for emissions methodology	See Chapter 4, Table 4.1		
Composting	See Chapter 4 for emissions methodology	See Chapter 4, Table 4.1		
Incineration and open burning	See Chapter 5 for emissions methodology	See Chapter 5		
Sources:				
¹ Based on expert judgment by Lead Authors of this section.				
² Emission factors calculated using default B ₀ and default MCF.				
³ See Annex 6A.4.				

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6.2.3.3 CHOICE OF ACTIVITY DATA

No refinement

6.2.3.4 TIME SERIES CONSISTENCY

No refinement

6.2.3.5 UNCERTAINTIES

No refinement

6.2.3.6 QA/QC, COMPLETENESS, REPORTING AND DOCUMENTATION

No refinement

6.3 NITROUS OXIDE EMISSIONS FROM DOMESTIC WASTEWATER

This section refines Section 6.3 of the *2006 IPCC Guidelines* by providing update and new guidance for estimating domestic N₂O emissions from WWTPs and provides new guidance on the estimation of N₂O emissions from wastewater after disposal of untreated wastewater or treated wastewater effluent into aquatic environments by accounting for losses of nitrogen prior to disposal.

6.3.1 Methodological issues

Nitrous oxide emissions can occur as emissions from WWTPs or as emissions from receiving aquatic environments following the disposal of untreated or treated wastewater effluent. This section describes how to estimate the N₂O produced during wastewater treatment, sludge treatment that occurs within the wastewater treatment system, and disposal of the wastewater. More recent research and field surveys have revealed that emissions in sewer networks and from nitrification or nitrification-denitrification processes at WWTPs, previously judged to be a minor source, may in fact result in more substantial emissions. N₂O is generated as a by-product of nitrification, or as an intermediate product of denitrification. There are many factors affecting N₂O emissions from wastewater treatment systems such as the temperature and dissolved oxygen concentration of the wastewater, and the specific operational conditions.

Emissions also result from untreated wastewater or wastewater treatment effluent that is discharged into aquatic environments. It is important to note that emissions are dependent on the nutrient-impacted status and oxygenation level of the aquatic environment receiving the wastewater discharge. The current methodology in the *2006 IPCC Guidelines* appropriately captures discharge to relatively clean and/or well-oxygenated environments. However, in the case of discharge to nutrient-impacted waters such as eutrophic lakes, estuaries and rivers, or locations where stagnant conditions occur, emissions can be significantly higher.

Three tiers of methods for estimating N₂O emissions from this category are summarised below:

The Tier 1 method applies default values for the emission factor and activity parameters. This method is considered *good practice* for countries with limited data.

The Tier 2 method follows the same method as Tier 1 but allows for incorporation of a country-specific emission factor and country-specific activity data. For example, a specific emission factor for a prominent treatment system based on field measurements could be incorporated under this method. Use of country-specific emission factor in the same region or climate area is also *good practice*. The amount of sludge removed for incineration, landfills, and agricultural land should be taken into consideration.

For a country with good data and advanced methodologies, a country-specific method could be applied as a Tier 3 method. For example, a more advanced country-specific method could be based on plant-specific emissions data from large wastewater treatment facilities or using country-specific measurements of nitrogen discharged to aquatic environments of varying nutrient-impacted status. Direct measurement methods would provide a more accurate measurement of N₂O production from each treatment plant.

6.3.1.1 CHOICE OF METHOD

A decision tree for domestic wastewater is included in new Figure 6.4.

The steps for *good practice* in inventory preparation for N₂O from domestic wastewater are as follows:

Step 1: Use new Equation 6.10 to estimate total nitrogen in wastewater and updated Equation 6.8 to estimate total nitrogen in wastewater effluent.

Step 2: Select the pathway and systems (see updated Figure 6.1) according to country activity data. Select the emission factor for each domestic wastewater treatment/discharge pathway or system.

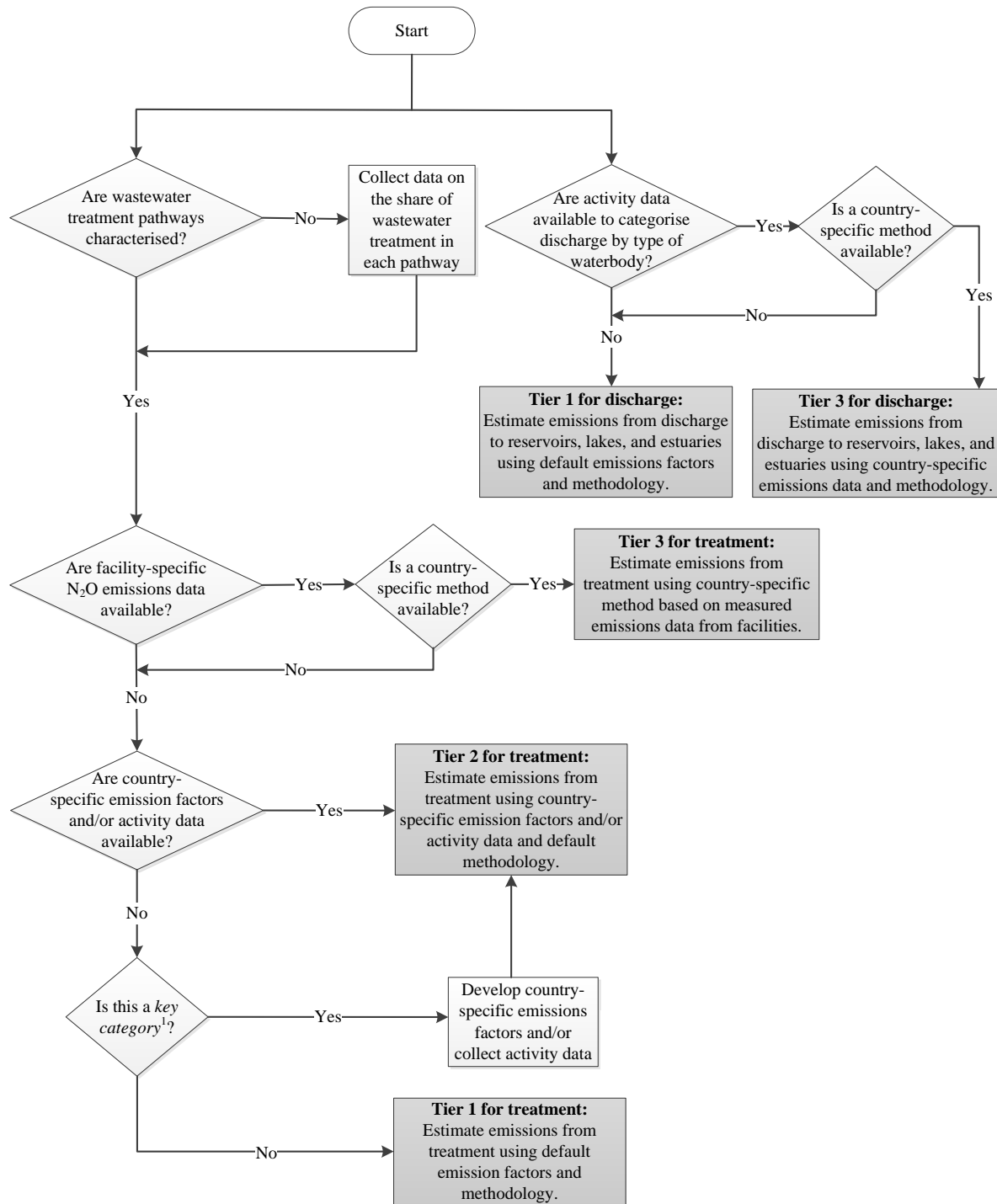
Step 3: Use updated Equation 6.9 to estimate emissions from wastewater treatment and sum the results for each treatment pathway/system.

Step 4: Use updated Equation 6.7 to estimate emissions from effluent, accounting for losses of nitrogen that occur within the wastewater treatment process including sludge removal and sum the results for each pathway/system. Emissions should also be calculated for nitrogen discharged as untreated wastewater.

As described earlier, the wastewater characterisation will determine the fraction of wastewater treated or disposed of by a particular system. To determine the use of each type of treatment or discharge system, it is *good practice* to refer to national statistics (e.g., from regulatory authorities). If these data are not available, wastewater associations or international organisations such as the World Health Organization (WHO) may have data on the system usage.

Otherwise, consultation with sanitation experts can help, and expert judgment can also be applied (see Chapter 2, Approaches to Data Collection, in Volume 1). Urbanisation statistics may provide a useful tool, e.g., city sizes and income distribution.

If activity data are available to categorize discharges to nutrient-impacted environments, these refinements introduce a new Tier 3 emission factor for those discharges.

Figure 6.4 (New) Decision tree for N₂O emissions from domestic wastewater

¹ See Volume 1 Chapter 4, "Methodological Choice and Identification of Key Categories" (noting Section 4.1.2 on limited resources), for discussion of key categories and use of decision trees.

EQUATION 6.9 (UPDATED)
N₂O EMISSIONS FROM DOMESTIC WASTEWATER TREATMENT PLANTS

$$N_2O\ Plants_{DOM} = \left[\sum_{i,j} \left(U_i \cdot T_{ij} \cdot EF_j \right) \right] \cdot TN_{DOM} \cdot \frac{44}{28}$$

Where:

848	$N_2O \text{ Plants}_{DOM}$	=	N_2O emissions from domestic wastewater treatment plants in inventory year, kg
849			N_2O/yr
850	TN_{DOM}	=	total nitrogen in domestic wastewater in inventory year, kg N/yr. See new Equation
851			6.10.
852	U_i	=	fraction of population in income group i in inventory year. See Table 6.5.
853	T_{ij}	=	degree of utilisation of treatment/discharge pathway or system j , for each income group
854			fraction i in inventory year. See Table 6.5.
855	i	=	income group: rural, urban high income and urban low income
856	j	=	each treatment/discharge pathway or system
857	EF_j	=	emission factor for treatment/discharge pathway or system j , kg N_2O -N/kg N
858	The factor 44/28 is for the conversion of kg N_2O -N into kg N_2O .		

859 It is also required to estimate direct N_2O emissions arising from the discharge of wastewater (treated or untreated)
 860 into aquatic receiving environments. The methodology for emissions from effluent is similar to that of N_2O
 861 emissions explained in Volume 4, Section 11.2.2 in Chapter 11 N_2O Emissions from Managed Soils, and CO_2
 862 Emissions from Lime and Urea Application. The simplified general equation is as follows:

EQUATION 6.7 (UPDATED)
 N_2O EMISSIONS FROM DOMESTIC WASTEWATER EFFLUENT

$$N_2O_{EFFLUENT,DOM} = N_{EFFLUENT,DOM} \cdot EF_{EFFLUENT} \cdot \frac{44}{28}$$

866 Where:

867	$N_2O_{EFFLUENT,DOM}$	=	N_2O emissions from domestic wastewater effluent in inventory year, kg N_2O/yr
868	$N_{EFFLUENT,DOM}$	=	nitrogen in the effluent discharged to aquatic environments, kg N/yr. See updated
869			Equation 6.8.
870	$EF_{EFFLUENT}$	=	emission factor for N_2O emissions from wastewater discharged to aquatic
871			systems, kg N_2O -N/kg N

872 The factor 44/28 is the conversion of kg N_2O -N into kg N_2O .

873 6.3.1.2 CHOICE OF EMISSION FACTORS

874 New Table 6.8a includes default Tier 1 EF values for use in estimation of N_2O emissions from waters receiving
 875 treated or untreated effluent (see Annex 6A.5). These refinements to the emission factors build on the current
 876 default value for $EF_{EFFLUENT}$ provided in Table 6.11 of the 2006 IPCC Guidelines. This existing emission factor is
 877 adequate for all discharges when using the Tier 1 methodology. However, it is expected to generate an
 878 underestimate of N_2O emissions in the case of discharge to nutrient-impacted waters such as eutrophic lakes,
 879 estuaries and rivers, or locations where stagnant conditions occur. Research published between 1978 and 2017
 880 provide data indicating that higher N_2O emissions occur when wastewater is discharged to nutrient-impacted
 881 (eutrophic) or hypoxic aquatic receiving environments. Accordingly, Table 6.8a provides a new $EF_{EFFLUENT}$ of
 882 0.019 g N_2O -N/g N (95% confidence limits 0.0041–0.091) for use with a Tier 3 method (see Annex 6A.6 for a
 883 discussion of information used to develop this new emission factor). The compiler can make use of many different
 884 indications of whether this higher EF value should be employed. Harmful algal blooms and hypoxia are the two
 885 most common symptoms of waterway nutrient over-enrichment (eutrophication). One possibility is to consider the
 886 dissolved oxygen status of the receiving waters. Waterway hypoxia is becoming increasingly common globally
 887 and definitions of hypoxia can vary depending on temperature, salinity, and the particular biota of interest.
 888 Dissolved oxygen concentrations of between 0.1–3.0 mg/L (or <30% of the oxygen saturation concentration) are
 889 typically classified as hypoxic (Vaquer-Sunyer & Duarte 2008; Naqvi *et al.* 2010; Rabalais *et al.* 2010) and are
 890 likely to result in increased N_2O yields from microbial metabolism of discharged wastewater nitrogen.

891 Oxygen status can be a highly dynamic indicator in practice, and compilers may have better access to information
 892 regarding the nutrient (trophic) status of receiving waters. Various thresholds for the boundary between
 893 mesotrophic and eutrophic conditions have been proposed, such as Secchi depth (suggestions range from 2 to 3.6
 894 m), total phosphorus (20–75 $\mu g/L$), total nitrogen (0.3–5 mg/L), benthic chlorophyll (70 mg/m²) and suspended
 895 chlorophyll concentration (5–30 $\mu g/L$) (Dodds *et al.* 1998; Burns *et al.* 2009; Mateo-Sagasta & Burke 2010).
 896 Compilers may also consider whether there have been reports of algal blooms, sulphurous odours, fish kills or

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other indications of eutrophic and/or hypoxic conditions in the discharge environment. Global datasets exist which may assist compilers in assessing the risk of nutrient impacts in different regions (Diaz *et al.* 2011; Xie & Ringler 2017). It is important to note that the context for application of this new Tier 3 EF_{EFFLUENT} factor is not necessarily the same as for nitrogen runoff from agriculture, due to the combination of reducing conditions and high organic loading in many wastewaters which are unlikely to exist in agricultural runoff systems.

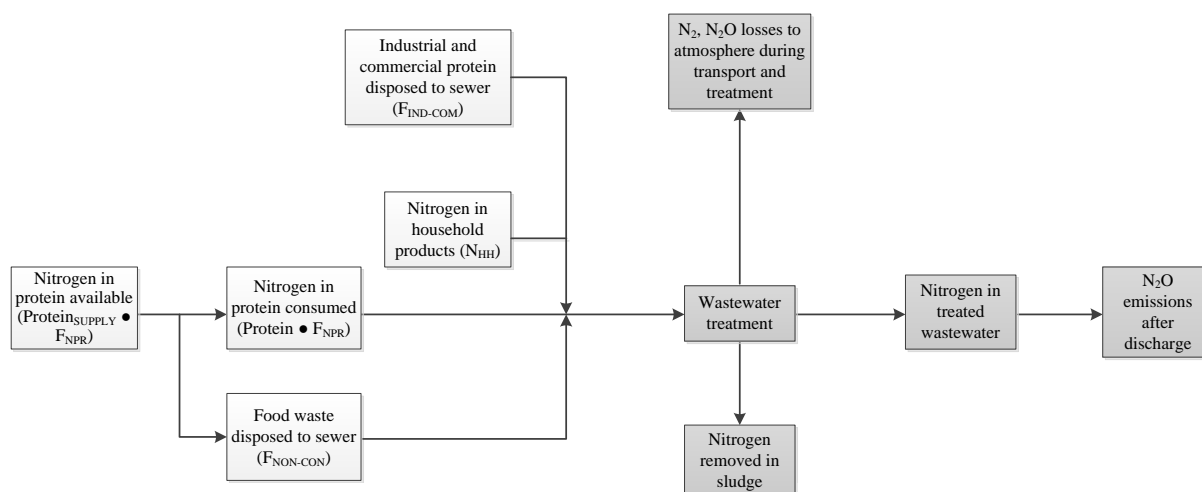
TABLE 6.8A (NEW) DEFAULT EF VALUES FOR DOMESTIC AND INDUSTRIAL WASTEWATER			
Type of treatment and discharge pathway or system	Comments	EF ¹ (kg N ₂ O-N/kg N)	Range
Discharge from treated or untreated system, EF _{EFFLUENT}			
Freshwater, estuarine, and marine discharge (Tier 1)	Based on limited field data and on specific assumptions regarding the occurrence of nitrification and denitrification in rivers and in estuaries	0.005 ²	0.0005 – 0.075
Nutrient-impacted and/or hypoxic freshwater, estuarine, and marine environments (Tier 3, if needed)	Higher emissions are associated with nutrient-impacted/hypoxic water such as eutrophic lakes, estuaries and rivers, or locations where stagnant conditions occur	0.019 ²	0.0041 – 0.091
Discharge to soil	Dry climate	0.005	0.0005 – 0.075
Discharge to soil	Wet climate	0.005	0.0005 – 0.075
Wastewater treatment system, EF _{plants}			
Centralised, aerobic treatment plant	N ₂ O is variable and can be significant	0.016 ¹	0.00016 – 0.045
Anaerobic reactor	N ₂ O is not significant	0	0 – 0.001
Anaerobic lagoons	N ₂ O is not significant	0	0 – 0.001
Constructed wetlands	See 2013 Supplement to the 2006 IPCC Guidelines for National Greenhouse Gas Inventories: Wetlands (IPCC 2014)		
Septic tank	N ₂ O is not significant	0	0 – 0.001
Septic tank + land dispersal field	N ₂ O is emitted by the soil dispersal system	0.0045	0 – 0.001
Latrine	N ₂ O is not significant	0	0 – 0.001
Sludge treatment system			
Anaerobic digester for sludge	N ₂ O is not significant	0	0
Composting	See Chapter 4 for emissions methodology	See Chapter 4, Table 4.1	
Incineration and open burning	See Chapter 5 for emissions methodology	See Chapter 5	
Sources: Based on scientific literature and expert judgment by Lead Authors of this section.			
¹ See Annex 6A.5.			
² See Annex 6A.6.			

6.3.1.3 CHOICE OF ACTIVITY DATA

The methodology to calculate the amount of N content in wastewater effluent ($N_{\text{EFFLUENT,DOM}}$) presented in Equation 6.8 of the 2006 IPCC Guidelines currently does not account for the removal of ammonia-N via nitrification-denitrification processes in the sewer and the WWTP, or the removal of N in sludge, which can result in a significant overestimate of N discharged to aquatic receiving environments. Typical total N concentration in raw urban sewage is about 40 mg/L (range 20–70 mg/L) (Tchobanoglous *et al.* 2014), whereas effluent treated in conventional activated sludge facilities (with nitrification) has a total of about 25 mg/L. Plants with biological nutrient removal (with denitrification) regularly achieve an effluent total N of 5 mg/L or less. The difference is due both to N removed in sludge (see Table 2.4A (New), Chapter 2 of Volume 5 regarding the N content of sewage sludge) versus N loss to the atmosphere (see new Figure 6.5).

This section updates the methodology presented in the 2006 IPCC Guidelines to estimate the amount of N content in wastewater entering treatment (TN_{DOM}), the loss or removal of N through the treatment process (either through biological conversion or removal with sludge) (N_{REM}), and the N content in wastewater discharged to aquatic systems ($N_{EFFLUENT,DOM}$).

Figure 6.5 (New) Nitrogen in domestic wastewater treatment



NITROGEN IN DOMESTIC WASTEWATER (TN_{DOM})

The activity data that are needed for estimating TN_{DOM} in domestic wastewater include the population associated with the wastewater, the average annual per capita protein consumed (kg/person/yr), and factors to account for N in non-consumed protein entering the wastewater and other N from household, industrial, and commercial sources co-discharged into the sewer system. The total N in wastewater for each treatment pathway is estimated as follows:

EQUATION 6.10 (NEW) TOTAL NITROGEN IN DOMESTIC WASTEWATER BY TREATMENT PATHWAY

$$TN_{DOM_k} = (P_{treatment_k} \bullet Protein \bullet F_{NPR} \bullet N_{HH} \bullet F_{NON-CON} \bullet F_{IND-COM})$$

Where:

- TN_{DOM_k} = total annual amount of nitrogen in domestic wastewater for treatment pathway k , kg N/yr
- $P_{treatment_k}$ = human population who are served by the treatment pathway k , person/yr
- $Protein$ = annual per capita protein consumption, kg protein/person/yr
- F_{NPR} = fraction of nitrogen in protein, default = 0.16 kg N/kg protein
- $F_{NON-CON}$ = factor for nitrogen in non-consumed protein disposed in sewer system, kg N/kg N. See new Table 6.10a.
- $F_{IND-COM}$ = factor for industrial and commercial co-discharged protein into the sewer system, kg N/kg N
- N_{HH} = additional nitrogen from household products added to the wastewater, default is 1.1 (some country data are in new Table 6.10a).

If national statistics on protein consumed or protein supply are not available, Food Balance Sheets of FAOSTAT can be used as activity data on per capita “protein supply quantity.” This information represents the total amount of protein available to the population but must be adjusted to reflect the fraction of protein consumed (FPC), according to the new Equation 6.10a.

EQUATION 6.10A (NEW) ESTIMATION OF PROTEIN CONSUMED

$$Protein = Protein_{SUPPLY} \bullet FPC$$

Where:

- $Protein_{SUPPLY}$ = annual per capita protein supply, kg protein/person/yr

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FPC = Fraction of protein consumed. Default regional values are listed in the new Table 6.10a

Food that is not consumed may be disposed to the sewer (e.g., as result of the use of food waste disposals in some countries) and a factor for non-consumed protein ($F_{\text{NON-CON}}$) should be used to reflect this additional N entering wastewater (see new Table 6.10a). If food waste is disposed with solid waste, it is assumed that no additional N is entering waste water and $F_{\text{NON-CON}} = 1$.

Bath and laundry water may contain household chemicals (detergents, shampoos, softeners, dishwashing agents, WC fresheners, cosmetics, etc.) and a factor for N from household products (N_{HH}) should be used. The default factor for N_{HH} is 1.1 (Henze *et al.* 2008; Tjandraatmadja *et al.* 2008), but additional regional factors are provided in Table 6.10a.

TABLE 6.10A (NEW)
DEFAULT FACTORS FOR DOMESTIC WASTEWATER

Region ¹	Protein consumed ² as fraction of protein supply	$F_{\text{NON-CON}}$ ² in case food waste is disposed in sewer	Additional N from households' chemicals
Europe	0.85	1.09	1.08
North America and Oceania	0.80	1.13	1.17 (USA) 1.07 (Australia)
Industrialised Asia	0.86	1.08	No data
Sub-Saharan Africa	0.98	1.01	No data
North Africa, West and central Asia	0.90	1.06	No data
South and Southeast Asia	0.96	1.02	1.13 (India)
Latin America	0.92	1.04	No data

¹ See Annex 6A.7 for a list of countries by region

² Based on FAO (2011)

TABLE 6.10B (NEW)
ESTIMATE ON USE OF FOOD WASTE DISPOSAL IN SEWER¹

Country	Share of households disposing food waste in sewer
USA	50%
Canada	10%
Australia	12%
New Zealand	30%
UK	5%
Ireland	1 – 2%
Italy	Supported
Czech Republic	<1% / not allowed
Denmark	Rare
Belgium	Restricted / not allowed
The Netherlands	Not allowed
Germany	Not allowed

¹ Based on EPA (2008) and expert judgment by Lead Authors

Wastewater from industrial or commercial sources that is discharged into the sewer may contain protein (e.g., from grocery stores and butchers). The default for this fraction is 1.25 for centralised treatment and 0 for decentralised treatment systems (septic system, latrines, discharge).

NITROGEN IN WASTEWATER EFFLUENT ($N_{\text{EFFLUENT},\text{DOM}}$)

The total nitrogen in wastewater effluent is estimated as follows:

EQUATION 6.8 (UPDATED)
TOTAL NITROGEN IN DOMESTIC WASTEWATER EFFLUENT

$$N_{\text{EFFLUENT},\text{DOM}} = \sum_k \left[(TN_{\text{DOM}} \bullet FWT_k) \bullet (1 - N_{\text{REM},k}) \right]$$

Where:

- $N_{\text{EFFLUENT},\text{DOM}}$ = total nitrogen in the wastewater effluent discharged to aquatic environments in inventory year, kg N/yr
- TN_{DOM} = total nitrogen in domestic wastewater in inventory year, kg N/yr. See new Equation 6.10.
- FWT_k = fraction of wastewater treated exclusively by each wastewater treatment type k in inventory year. See Table 6.5.
- k = each wastewater treatment type used in inventory year
- N_{REM} = fraction of total wastewater nitrogen removed during wastewater treatment per treatment type k . See new Table 6.10c. Pathways for N removal include transfer to sludge and nitrification–denitrification with concomitant N loss to the atmosphere.

TABLE 6.10c (NEW)
WASTEWATER TREATMENT NITROGEN REMOVAL FRACTIONS (N_{REM})
ACCORDING TO TREATMENT TYPE

Treatment Type	Default	Range
No treatment	0 ¹⁻³	0 ¹⁻³
Primary (mechanical)	0.10 ¹⁻³	0.05 – 0.20 ¹⁻³
Secondary (biological)	0.40 ¹⁻³	0.35 – 0.55 ¹⁻³
Tertiary (advanced biological)	0.80 ¹⁻⁴	0.45 – 0.85 ¹⁻⁴
Septic tank	0.15 ¹⁻³	0.10 – 0.25 ¹⁻³
Septic tank + land dispersal field	0.68 ⁵	0.62 – 0.73 ⁵
Latrine	0.12 ⁶	0.07 – 0.21 ⁶
Sources: ¹ Kristensen <i>et al.</i> (2004) ² Van Drecht <i>et al.</i> (2009) ³ Based on expert judgment by Lead Authors of this section. ⁴ Ekama and Wentzel (2008) ⁵ Andreoli <i>et al.</i> (1979) ⁶ EMEP/EAA (2016)		

6.3.2 Time series consistency

No refinement

6.3.3 Uncertainties

Large uncertainties are associated with the IPCC default factors for N₂O. Updated Table 6.11 below includes uncertainty ranges based on expert judgment.

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TABLE 6.11 (UPDATED) N ₂ O METHODOLOGY DEFAULT DATA			
	Definition	Default Value	Range
Emission Factor			
EF	N ₂ O emission factor	See Table 6.8a	See Table 6.8a
Activity Data			
P	Number of people in country	Country-specific	± 10 %
Protein	Annual per capita protein consumption	Country-specific	± 10 %
F _{NPR}	Fraction of nitrogen in protein (kg N/kg protein)	0.16	0.15 – 0.17
T _{plant}	Degree of utilisation of large WWT plants	Country-specific	± 20 %
F _{NON-CON}	Factor to adjust for non-consumed protein, based on available protein	1.0 for countries with no in-sink disposals, 1.1 for countries with in-sink disposals	0.9 – 1.2
	Factor to adjust for non-consumed protein, based on consumed protein	1.1 for countries with no in-sink disposals, 1.25 for countries with in-sink disposals	1.0 – 1.4
F _{IND-COM}	Factor to allow for co-discharge of industrial nitrogen into sewers. For countries with significant fish processing plants, this factor may be higher. Expert judgment is recommended.	1.25	1.0 – 1.5
N _{REM}	Factor to account for losses of nitrogen prior to discharge	See Table 6.10c	See Table 6.10c

6.3.4 QA/QC, Completeness, Reporting and Documentation

No refinement

6.4 NITROUS OXIDE EMISSIONS FROM INDUSTRIAL WASTEWATER

6.4.1 Methodological issues

This section refines the *2006 IPCC Guidelines* by adding new guidance for estimating N₂O emissions from industrial WWTPs and refines the estimation of N₂O emissions from wastewater after disposal of untreated wastewater or wastewater treatment effluent into aquatic environments by accounting for losses of nitrogen prior to disposal.

6.4.1.1 CHOICE OF METHOD

A decision tree for industrial wastewater is included in new Figure 6.6.

The steps for *good practice* in inventory preparation for N₂O from industrial wastewater are as follows:

Step 1: Use new Equation 6.13 to estimate total nitrogen in wastewater.

Step 2: Select the pathway and systems (See updated Figure 6.1) according to country activity data. Select the emission factor for each domestic wastewater treatment/discharge pathway or system.

Step 3: Use new Equation 6.11 to estimate emissions from wastewater treatment and sum the results for each pathway/system.

Step 4: Use new Equation 6.12 to estimate emissions from effluent, accounting for losses of nitrogen that occur within the wastewater treatment process, including sludge removal, and sum the results for each pathway/system.

As described earlier, the wastewater characterisation will determine the fraction of wastewater treated or disposed of by a particular system. To determine the use of each type of treatment or discharge system, it is *good practice* to refer to national statistics (e.g., from regulatory authorities). If these data are not available, industry associations may have data on the system usage. Otherwise, consultation with industry experts can help, and expert judgment can also be applied (see Chapter 2 Approaches to Data Collection, Volume 1).

EQUATION 6.11 (NEW)
N₂O EMISSIONS FROM INDUSTRIAL WASTEWATER TREATMENT PLANTS

$$N_2O\ Plants_{IND} = \left[\sum_i (T_{i,j} \bullet EF_j \bullet TN_{IND_i}) \right] \times \frac{44}{28}$$

Where:

N₂O Plants_{IND} = N₂O emissions from industrial wastewater treatment plants in inventory year, kg N₂O/yr

TN_{INDi} = total nitrogen in wastewater from industry *i* in inventory year, kg N/yr. See new Equation 6.13.

T_{ij} = degree of utilisation of treatment/discharge pathway or system *j*, for each industry *i* in inventory year

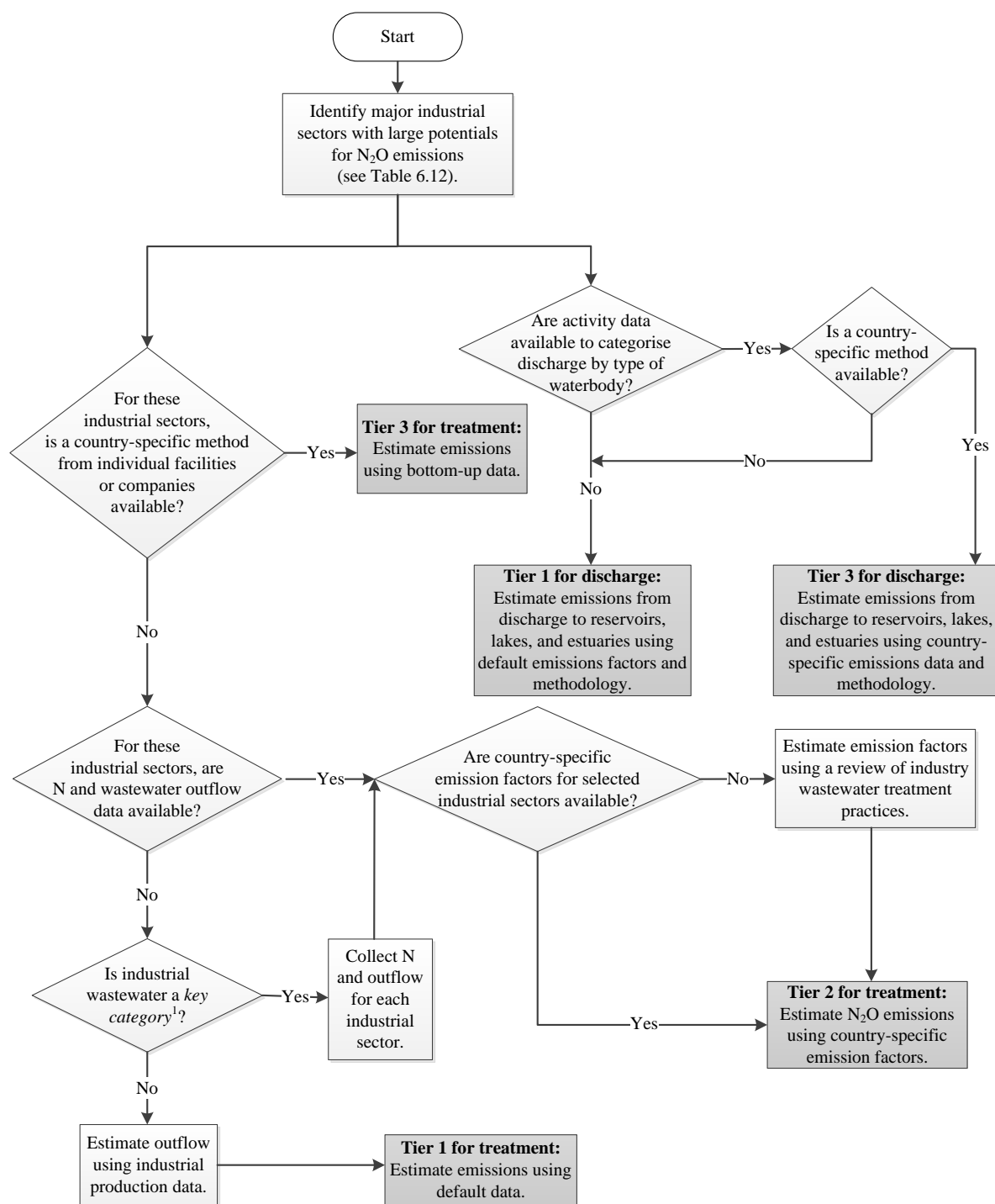
i = industry

j = each treatment/discharge pathway or system

EF_{*j*} = emission factor for treatment/discharge pathway or system *j*, kg N₂O-N/kg N

The factor 44/28 is for the conversion of kg N₂O-N into kg N₂O.

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Figure 6.6 (New) Decision tree for N₂O emissions from industrial wastewater

¹ See Volume 1 Chapter 4, "Methodological Choice and Identification of Key Categories" (noting Section 4.1.2 on limited resources), for discussion of key categories and use of decision trees.

It is also required to estimate N₂O emissions from wastewater treatment effluent that is discharged into aquatic environments. The methodology for emissions from effluent is like that of N₂O emissions explained in Volume 4, Section 11.2.2, Chapter 11, N₂O Emissions from Managed Soils, and CO₂ Emissions from Lime and Urea Application. The simplified general equation is as follows:

EQUATION 6.12 (NEW)
N₂O EMISSIONS FROM INDUSTRIAL WASTEWATER EFFLUENT

$$N_2O_{Effluent_{IND}} = N_{EFFLUENT,IND} \times EF_{EFFLUENT} \times \frac{44}{28}$$

Where:

$N_2O_{Effluent_{IND}}$ = N₂O emissions from industrial wastewater effluent in inventory year, kg N₂O/yr

$N_{EFFLUENT,IND}$ = nitrogen in the industrial wastewater effluent discharged to aquatic environments, kg N/yr. See new Equation 6.14.

$EF_{EFFLUENT}$ = emission factor for N₂O emissions from wastewater discharged to aquatic systems, kg N₂O-N/kg N

The factor 44/28 is for the conversion of kg N₂O-N into kg N₂O.

6.4.1.2 CHOICE OF EMISSION FACTORS

New Table 6.8a includes default EF values for N₂O emissions.

6.4.1.3 CHOICE OF ACTIVITY DATA

TOTAL NITROGEN IN INDUSTRIAL WASTEWATER (TN_{IND})

The activity data for this source category is the amount of total nitrogen (TN) in the industrial wastewater entering treatment (TN_{IND}). This parameter is a function of industrial output (product) P (tonnes/yr), wastewater generation W (m³/ton of product), and total N concentration in the untreated wastewater (kg TN/m³). See new Equation 6.13. The following steps are required for determination of TN_{IND}:

- (i) Identify the industrial sectors that generate wastewater with large quantities of N, by evaluating total industrial product, N in the wastewater, and wastewater produced.
- (ii) Identify industrial sectors that use treatment systems with N₂O emissions factors (see new Table 6.8a). Experience has shown that usually three or four industrial sectors are *key*.

For each selected sector estimate total N in the industrial wastewater (TN_{IND}).

EQUATION 6.13 (NEW)
TOTAL NITROGEN IN INDUSTRIAL WASTEWATER

$$TN_{IND_i} = P_i \bullet W_i \bullet TN_i$$

Where:

TN_{IND_i} = total nitrogen in wastewater entering treatment for industry *i*, kg TN/yr

i = industrial sector

P_i = total industrial product for industrial sector *i*, t/yr

W_i = wastewater generated for industrial sector *i*, m³/t_{product}

TN_i = total nitrogen in untreated wastewater for industrial sector *i*, kg TN/m³

Industrial production data and wastewater generation rates may be obtained from national statistics, regulatory agencies, wastewater treatment associations or industry associations. In some cases, quantification of the N loading in the wastewater may require expert judgment. In some countries, N content and total water usage per sector data may be available directly from a regulatory agency. New Table 6.12 provides examples that could be used as default values for industries that may be considered key sources of N₂O. These values should be used with caution, because they are industry-, process-, and country-specific.

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TABLE 6.12 (NEW)
EXAMPLES OF INDUSTRIAL WASTEWATER DATA

Industry Type	Wastewater Generation W (m ³ /tonne)	Range for W (m ³ /tonne)	Total Nitrogen (TN) (kg/m ³)	TN Range (kg/m ³)
Alcohol refining	24 ²	16 – 32 ²	2.4 ²	0.94 – 3.86 ²
Beer & malt	6.3 ²	5.0 – 9.0 ²	0.055 ³	0.025 – 0.08 ³
Fish processing	5 ²	2 – 8 ²	0.60 ²	0.21 – 0.98 ²
Iron and steel manufacturing	5 ¹	0.004 – 10.4 ⁴	0.25 ¹	0.0004 – 0.524 ⁴
Meat & poultry	13 ²	8 – 18 ²	0.19 ²	0.17 – 0.20 ²
Nitrogen fertiliser	2.89 ²	0.46 – 8.3 ²	0.5 ²	0.1 – 0.8 ²
Plastics & resins	0.6 ⁵	0.3 – 1.2 ⁵	0.25 ⁶	No range provided
Starch production	9 ²	4 – 18 ²	0.9 ²	0.8 – 1.10 ²

¹ Based on expert judgment by Lead Authors of this section.
² IPCC (2014)
³ Simate *et al.* (2011)
⁴ US EPA (2002a)
⁵ Doorn *et al.* (1997)
⁶ Li *et al.* (2016)

NITROGEN REMOVED FROM WASTEWATER (N_{REM})

Nitrogen removal by different treatment facilities can range from 10–85 percent. Default values for the fraction of nitrogen removed by type of wastewater treatment system are presented in Table 6.10c.

NITROGEN IN WASTEWATER EFFLUENT (N_{EFFLUENT,IND})

The total nitrogen in the industrial wastewater effluent is estimated as follows:

EQUATION 6.14 (NEW)
TOTAL NITROGEN IN INDUSTRIAL WASTEWATER EFFLUENT

$$N_{EFFLUENT,IND} = \sum_k [TN_{IND,i} \times FWT_k \times (1 - N_{REM,k})]$$

Where:

- N_{EFFLUENT,IND} = total annual amount of nitrogen in the industrial wastewater effluent, kg N/yr
- TN_{IND,i} = total nitrogen in wastewater entering treatment for industry *i*, kg TN/yr
- FWT_k = fraction of wastewater treated exclusively by each wastewater treatment type *k* in inventory year. See Table 6.5.
- k = each wastewater treatment type used in inventory year
- N_{REM,k} = fraction of total wastewater nitrogen removed during wastewater treatment per treatment type *k*. See new Table 6.10c.

6.4.2 Time series consistency

Once an industrial sector is included in the inventory calculation, it should be included for each subsequent year. If the inventory compiler adds a new industrial sector to the calculation, then he or she should re-calculate the entire time series so that the method is consistent from year to year. General guidance on recalculation of estimates through time series is provided in Volume 1, Chapter 5, Time Series Consistency.

If a country decides to incorporate plant emissions into the estimate, this change must be made for the entire time series. Potential sludge removal should be treated consistently across years in the time series.

6.4.3 Uncertainties

Uncertainty estimates for EF, P, W and TN are provided in new Table 6.13. The estimates are based on expert judgment.

TABLE 6.13 (NEW) DEFAULT UNCERTAINTY RANGES FOR INDUSTRIAL WASTEWATER	
Parameter	Uncertainty Range
Emission Factor	
N ₂ O emission factor (EF)	± 30%
Activity Data	
Industrial production (P)	± 25% Use expert judgment regarding the quality of data source to assign more accurate uncertainty range.
Wastewater/unit production (W)	These data can be very uncertain as the same sector might use different waste handling procedures at different plants and in different countries. The product of the parameters (W•TN) is expected to have less uncertainty. An uncertainty value can be attributed directly to TN concentration. –50 %, +100% is suggested (i.e., a factor of 2).
TN/unit wastewater (TN)	
Source: Expert Judgment by Lead Authors of this section.	

6.4.4 QA/QC, Completeness, Reporting and Documentation

QA/QC

It is *good practice* to conduct quality control checks and quality assurance procedures as outlined in Chapter 6, QA/QC and Verification, of Volume 1. Below, some fundamental QA/QC procedures include:

- For industrial wastewater, inventory compilers may review the secondary data sets (e.g., from national statistics, regulatory agencies, wastewater treatment associations or industry associations) that are used to estimate and rank industrial TN waste output. Some countries may have regulatory control over industrial discharges, in which cases significant QA/QC protocols may already be in place for the development of the wastewater characteristics on an industry basis.
- For industrial wastewater, inventory compilers should cross-check values for EFs against those from other national inventories with similar wastewater characteristics.
- If sludge removal is reported in the wastewater inventory, check for consistency with the estimates for sludge applied to agriculture soils, sludge incinerated, and sludge deposited in solid waste disposal.
- For countries that use country-specific parameters or higher tier methods, inventory compilers should cross-check the national estimates with emissions using the IPCC default method and parameters.

COMPLETENESS

Completeness for estimating emissions from industrial wastewater depends on an accurate characterisation of industrial sectors that produce nitrogen-laden wastewater. In most countries, approximately 3–4 industrial sectors will account for the majority of the nitrogenous wastewater volume, so the inventory compilers should ensure that these sectors are covered. Periodically, the inventory compilers should re-survey industrial sources, particularly if some industries are growing rapidly.

This category should only cover industrial wastewater treated onsite. Emissions from industrial wastewater released into domestic sewer systems should be addressed and included with domestic wastewater.

Some sludge from industrial wastewater treatment may be incinerated or deposited in landfills or on agricultural lands. This constitutes an amount of N that should be subtracted from effluent N. It is *good practice* to be consistent across sectors: the amount of N that is removed as sludge should be equal to the amount of sludge disposed at landfills, applied to agricultural soils, incinerated, or treated elsewhere.

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It is *good practice* to document and report a summary of the methods used, activity data and emission factors. When country-specific methods and/or emission factors are used, the reasoning for the choices as well as references to how the country-specific data (measurements, literature, expert judgment, etc.) have been derived (measurements, literature, expert judgment, etc.) should be documented and included in the reporting.

If sludge is incinerated, landfilled, or spread on agricultural lands, the quantities of sludge and associated emissions should be reported in the waste incineration, solid waste disposal systems, or agricultural categories, respectively.

More information on reporting and documentation can be found in Volume 1, Chapter 6, Section 6.11 Documentation, archiving and reporting.

Appendix 6A.1 Abiogenic (fossil) CO₂ emissions from wastewater treatment and discharge: Basis for Future Methodological Development

The 2006 IPCC Guidelines assume that organic carbon present in wastewater derives from modern (biogenic) organic matter in human excreta or food waste (Doorn *et al.* 2006). Consequently, CO₂ emissions from wastewater treatment according to those Guidelines are also considered wholly biogenic and are discounted from international greenhouse gas accounting inventories, since they do not represent a transfer of carbon from the lithosphere to the atmosphere. In comparison, fossil organic carbon (that with turnover timescales exceeding 10⁶ years) is considered to have a role in climate change and is accounted for in international greenhouse gas emissions inventories. In the 2006 IPCC Guidelines, it is assumed that emissions from biogas flaring are not significant because the CO₂ emissions are of biogenic origin and the CH₄ and N₂O emissions are very small so *good practice* in the Waste Sector does not require their estimation. If countries wish to report such emissions, they should be reported under the Waste Sector. A discussion of emissions from flares and more detailed information are given in Volume 2, Energy, Chapter 4.2. However, the presence of fossil organic carbon in sewage also implies the emission of additional fossil CO₂ from wastewater treatment facilities, sludge management, and environmental recipients of treated or untreated wastewater.

Data emerging since the 2006 IPCC Guidelines indicate that wastewater contains an appreciable amount of abiogenic (fossil) organic carbon, with this fossil carbon thought to be derived from the use of petroleum-based products (domestically and commercially). These products include: cosmetics; pharmaceuticals; surfactants; detergents and food additives (Law *et al.* 2013). Additionally, direct dosing of synthetic, fossil-derived organic substrates (e.g., methanol) can occur at wastewater treatment plants to enhance denitrification performance (Schneider *et al.* 2015). Despite early indications of the potential for significant fossil organic carbon fractions in sewage sludge (Turekian & Benoit 1981), fossil carbon in wastewater was first detailed by Griffith *et al.* (2009) following the sampling of treated effluent from 12 predominantly domestic wastewater treatment plants within the Hudson and Connecticut River watersheds, USA. Since then, several studies have surveyed fossil organic carbon in wastewater from Japan (Nara *et al.* 2010; Toyoda *et al.* 2011), Australia (Law *et al.* 2013), Denmark (Yoshida *et al.* 2014) and North America (Schneider *et al.* 2015).

Research published since the 2006 IPCC Guidelines (Griffith *et al.* 2009; Law *et al.* 2013; Yoshida *et al.* 2014; Schneider *et al.* 2015; Tseng *et al.* 2016) gives an emerging consensus figure for the fossil wastewater organic carbon fraction at some 4–14%; although more recent work suggests this can be as high as 28% (Tseng *et al.*, 2016) or 51% (Nara *et al.* 2010; recalculated by Tseng *et al.* 2016) in some influents. Schneider *et al.* (2015) put the fossil content of activated sludge process off-gases at 10–15%. Law *et al.* (2013) put the subsequent fossil carbon fraction in anaerobic digester biogas CO₂ at 2.1±0.2% due to greater recalcitrance of fossil carbon during anaerobic digestion, and total additional scope 1 wastewater treatment plant emission load from previously unaccounted fossil carbon somewhere between 2–12%. Tseng *et al.* (2016) put this figure at some 13 to 24% higher without and with energy recovery respectively.

Based on the above data, countries are encouraged to evaluate if such emissions should be reported, particularly those countries that have higher levels of fossil carbon in wastewater. In addition, future improvements to the IPCC Guidelines should include a method for estimating these non-biogenic emissions associated with wastewater treatment operations and wastewater discharges.

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TABLE 6Ap.1 (NEW)
SUMMARY OF LITERATURE INVESTIGATING FOSSIL ORGANIC CARBON IN WASTEWATER

Wastewater treatment system	Detection/measurement approach	Fossil C fraction in various wastewater streams	Reference	Comments
Effluent from 12 WWTPs (using conventional activated sludge) sampled, USA	Effluent grab samples; ¹⁴ C analysed via isotope ratio mass spectrometry	25% of treated wastewater dissolved organic carbon (DOC) and 14% of treated wastewater particulate organic carbon (POC)	Griffith <i>et al.</i> (2009)	WWTPs handle predominantly domestic wastewater. Secondary treated effluent sampled only (i.e. not raw influent), so results may overestimate true fossil carbon fraction in raw sewage, since heterotrophic bacteria are known to preferentially utilize young (¹⁴ C-enriched) carbon for assimilation (Raymond & Bauer 2001). Average reported effluent DOC concentration was 8.7 mg/L.
Four large activated sludge WWTPs in Brisbane, Australia (two receiving domestic sewage and two receiving 15% industrial load)	Range of sampling locations (influent, primary- and secondary-treated effluent, waste activated sludge (WAS), digested biosolids)	Fossil fraction of total organic carbon (TOC) in domestic WWTP influent 4–7% and 8–14% for WWTPs receiving 15% industrial load (5–14% overall range); ≈29–50% of this influent fossil C is transformed to CO ₂ during secondary activated sludge treatment (1.4–6.3% of influent TOC). Higher fossil C fraction in WAS from WWTPs receiving industrial load (8–14%) versus domestic WWTPs (6–7%). 2.1(±0.2)% of biogas CO ₂ is fossil in origin.	Law <i>et al.</i> (2013)	Higher fossil C fractions in WWTPs receiving industrial wastewater load. Equivalent concentrations of fossil organic carbon in raw wastewater were between 6–35 mg/L. Majority of fossil organic carbon in raw wastewater is present in particulate form (>80%; 5 to 29 mg/L), whereas dissolved fossil carbon levels are relatively small (1–6 mg/L).
Influent to Avedøre WWTP, Denmark	Single 24-hour flow proportionate composite influent sample collected in February 2013. Radiocarbon isotope ratio method (ASTM-D6866-12).	14(±3)% in influent wastewater	Yoshida <i>et al.</i> (2014)	Avedøre WWTP receives 15% industrial load from adjacent pharmaceuticals industry. Figure of 14% is corrected for cellulosic biogenic carbon from toilet paper which may contain elevated levels of ¹⁴ C due to historical atomic bomb detonations and underestimate fossil C fraction.
Modified Ludzack–Ettinger (MLE) activated sludge process with biological nitrification–denitrification, USA	¹⁴ C content of emitted CO ₂ measured twice a day for five days in early spring using floating chambers	11.4–15.1% (mean 12.83%) based on measured CO ₂ emissions from secondary treatment reactor	Schneider <i>et al.</i> (2015)	
Three municipal WWTPs and waste stabilisation ponds, two industrial WWTPs	Raw and partially treated wastewater, gas and sludge samples taken during 2010–2013	2–28% in the primary influent	Tseng <i>et al.</i> (2016)	Article provides a tabulation of results from 6 separate research papers.

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TABLE 6Ap.1 (NEW) (CONTINUED) SUMMARY OF LITERATURE INVESTIGATING FOSSIL ORGANIC CARBON IN WASTEWATER				
Wastewater treatment system	Detection/measurement approach	Fossil C fraction in various wastewater streams	Reference	Comments
Assumed conventional activated sludge-based WWTP with primary gravity sedimentation, Japan	Fresh domestic sewage, primary-treated and secondary-treated wastewater sampled	$\Delta^{14}\text{C}$ values (‰) of sewage DOC was more negative (^{14}C ages in the order of ≈ 1000 – 5000 years) than nearby lake and river water; no fossil C fraction given.	Nara <i>et al.</i> (2010)	No information on relative fossil C fraction given.
A2O WWTP, Japan	13 wastewater samples collected along the treatment train during March, 2008	$\delta^{13}\text{C}$ signature of -50.7‰ (depleted with respect to background air)	Toyoda <i>et al.</i> (2011)	WWTP received municipal wastewater

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Annex 6A.1 Summary data for pit latrine use, no sanitation facility, and groundwater use by country

TABLE 6A.1 (NEW)					
SUMMARY DATA FOR PIT LATRINE USE, NO SANITATION FACILITY, AND GROUNDWATER USE BY COUNTRY					
Country	Data Source ¹	Report Year	Pit Latrine Use for Sanitation (%)	No Sanitation Facility (%)	Groundwater Use for Drinking (%)
Afghanistan	MICS	2010–2011	42.5	17.7	62.9
Albania	DHS	2008–2009	24.0	0.0	20.8
Algeria	MICS	2006	2.2	4.9	13.4
Angola	MICS	2001	31.0	38.7	37.5
Armenia	DHS	2010	30.4	0.0	2.5
Azerbaijan	DHS	2006	58.8	0.3	25.7
Bangladesh	DHS	2007	60.1	7.5	90.8
Belarus	MICS	2005	27.5	0.0	12.5
Belize	MICS	2006	40.7	2.1	7.5
Benin	DHS	2006	11.7	69.5	38.6
Bhutan	MICS	2010	52.7	3.4	1.5
Bolivia	DHS	2008	25.7	28.4	7.9
Bosnia and Herzegovina	MICS	2006	2.5	0.1	15.7
Botswana	MICS	2000	57.2	16.2	2.9
Brazil	DHS	1996	42.2		21.3
Burkina Faso	MICS	2006	1.0	62.6	68.9
Burundi	MICS	2005	92.7	3.0	69.0
Cambodia	DHS	2010	5.7	55.3	50.0
Cameroon	MICS	2006	82.6	7.5	49.9
Central African Republic	MICS	2006	75.4	22.3	69.1
Chad	DHS	2004	24.3	74.1	77.2
China	CHS04	2004	49.9	2.3	20.1
Colombia	DHS	2010	0.7	4.8	3.2
Comoros	MICS	2000	95.0	0.7	7.9
Congo	DHS	2005	84.4	10.2	30.6
Côte d'Ivoire	MICS	2006	42.7	34.0	51.8
Cuba	MICS	2010–2011	25.7	1.0	18.2
Democratic People's Republic of Korea	MICS	2009	37.4	0.0	10.8
Democratic Republic of the Congo	MICS	2010	80.0	14.5	59.1
Djibouti	MICS	2006	73.4	3.9	2.7
Dominican Republic	DHS	2007	47.6	36.0	9.7
Egypt	DHS	2008		0.4	4.0
Eritrea	DHS	2002	9.0	74.3	45.2

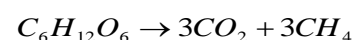
TABLE 6A.1 (NEW) (CONTINUED)					
SUMMARY DATA FOR PIT LATRINE USE, NO SANITATION FACILITY, AND GROUNDWATER USE BY COUNTRY					
Country	Data Source ¹	Report Year	Pit Latrine Use for Sanitation (%)	No Sanitation Facility (%)	Groundwater Use for Drinking (%)
Ethiopia	DHS	2011	56.0	38.2	50.6
Gabon	DHS	2000	92.8	3.0	40.0
Gambia	MICS	2005–2006	80.6	4.4	41.0
Georgia	MICS	2005	57.8	0.0	25.7
Ghana	MICS	2010–2011	56.6	0.5	0.1
Guatemala	DHS	1998–1999	40.9	13.2	14.9
Guinea	DHS	2005	67.2	30.3	60.7
Guinea-Bissau	MICS	2006	4.0	31.1	79.1
Guyana	DHS	2009	43.3	1.0	3.5
Haiti	DHS	2005–2006	32.0	0.0	45.0
Honduras	DHS	2005–2006	39.8	16.7	15.8
India	DHS	2005–2006	12.9	56.2	58.6
Indonesia	DHS	2007	3.8	8.1	52.3
Iraq	MICS	2006	28.9	2.5	3.4
Jamaica	MICS	2005	77.2	0.6	3.3
Jordan	DHS	2009	43.5	0.0	0.0
Kazakhstan	MICS	2006	62.3	0.0	22.5
Kenya	DHS	2008–2009	67.3	14.5	41.4
Kyrgyzstan	MICS	2005–2006	82.0	0.1	10.1
Lao People's Democratic Republic	MICS	2006	31.7	50.1	48.7
Lesotho	DHS	2009	66.8	35.6	36.2
Liberia	DHS	2007	20.1	54.7	76.2
Madagascar	DHS	2008–2009	35.1	43.7	53.3
Malawi	DHS	2010	84.9	9.9	75.3
Maldives	DHS	2009	27.8	1.0	1.3
Mali	DHS	2006	60.2	19.6	69.9
Mauritania	MICS	2007	35.1	45.5	37.7
Mongolia	MICS	2005	67.1	13.4	60.7
Montenegro	MICS	2005–2006	7.7	0.3	9.2
Morocco	DHS	2003–2004	1.7	15.9	16.3
Mozambique	MICS	2008	52.7	41.8	55.9
Myanmar	MICS	2009–2010	74.9	7.0	73.6
Namibia	DHS	2006–2007	11.7	53.4	16.6
Nepal	DHS	2011	21.0	38.4	46.9
Nicaragua	DHS	2001	59.1	13.9	25.3
Niger	DHS	2006	21.5	78.0	74.3
Nigeria	MICS	2007	58.9	27.7	47.6
Pakistan	DHS	2006–2007	13.7	28.4	55.6

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TABLE 6A.1 (NEW) (CONTINUED) SUMMARY DATA FOR PIT LATRINE USE, NO SANITATION FACILITY, AND GROUNDWATER USE BY COUNTRY					
Country	Data Source¹	Report Year	Pit Latrine Use for Sanitation (%)	No Sanitation Facility (%)	Groundwater Use for Drinking (%)
Peru	DHS	2011	1.9	12.0	6.6
Philippines	DHS	2008	11.7	9.6	39.0
Republic of Moldova	DHS	2005	62.5	0.0	56.0
Rwanda	DHS	2010	96.6	1.1	59.0
Samoa	DHS	2009	10.0	0.1	3.8
Sao Tome and Principe	DHS	2008–2009	23.1	57.7	5.9
Senegal	DHS	2010–2011	57.1	16.5	27.6
Serbia	MICS	2010	4.5	0.0	8.5
Sierra Leone	MICS	2010	63.4	28.9	54.1
Somalia	MICS	2006	37.2	53.8	25.6
South Africa	DHS	2003	36.7	8.1	3.8
Sudan	MICS	2000	55.2	32.4	40.7
Suriname	MICS	2006	19.6	6.3	3.1
Swaziland	MICS	2010	69.7	15.4	19.3
Syrian Arab Republic	MICS	2006	18.2	1.0	7.7
Tajikistan	MICS	2005	85.5	0.4	14.2
TFYR Macedonia	MICS	2005	6.9	3.1	7.0
Thailand	MICS	2005–2006	1.4	0.8	12.3
Timor-Leste	DHS	2009–2010	28.0	35.8	48.7
Togo	MICS	2010	31.5	55.7	55.2
Trinidad and Tobago	MICS	2006	15.0	0.1	1.2
Turkey	DHS	2003	22.8	0.5	7.3
Turkmenistan	DHS	2000	71.3	0.6	22.9
Uganda	DHS	2006	66.4	11.8	73.5
Ukraine	DHS	2007	47.2	0.0	28.0
United Republic of Tanzania	DHS	2010	78.8	15.9	48.3
Uzbekistan	MICS	2006	87.4	0.0	20.1
Vanuatu	MICS	2007	77.3	3.2	22.7
Venezuela	MICS	2000	6.7	4.4	2.1
Viet Nam	MICS	2010–2011	18.2	6.4	43.7
Yemen	MICS	2006	42.1	21.4	35.9
Zambia	DHS	2007	57.1	23.5	47.1
Zimbabwe	DHS	2010–2011	42.6	28.3	64.0
Sources: ¹ MICS: Multiple Indicator Cluster Surveys, UNICEF (http://www.childinfo.org/mics.html); DHS: Demographic and Health Surveys, USAID (http://www.measuredhs.com); CHS04: Economic, Population, Nutrition, and Health Survey, data accessed from WHO/UNICEF Joint Monitoring Programme reports (http://wssinfo.org)					

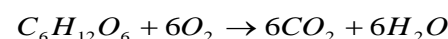
Annex 6A.2 Derivation of the maximum CH₄ producing potential (B₀) for domestic wastewater

The maximum CH₄ producing potential (B₀) for domestic wastewater is calculated theoretically by the anaerobic decomposition of glucose considering the total degradation of organic matter:



One mole of glucose weighs 180 g and produces 3 moles of methane which weighs $3 \times 16 = 48$ g. Therefore, the methane production rate per gram of glucose is $48 \text{ g} / 180 \text{ g} = 0.27 \text{ g methane} / \text{g glucose}$.

The complete oxidation of one mole of glucose (180 g) requires six moles of oxygen ($6 \times 32 \text{ g} = 192 \text{ g}$)



Then, the oxygen demand rate per gram of glucose is $192 \text{ g} / 180 \text{ g} = 1.067 \text{ g oxygen} / \text{g glucose}$

Finally, the maximum CH₄ producing potential (B₀) for domestic wastewater based of COD content is $48 \text{ g CH}_4 / 192 \text{ g oxygen} = 0.25 \text{ kg CH}_4 / \text{kg COD}$.

The COD/BOD ratio of a specific wastewater indicates the amount of organic matter difficult to degrade. For domestic wastewater, a typical COD/BOD ratio is 2.4 based on empirical tests. However, the ratio between the components in any given domestic wastewater stream may vary due to contributions from other sources, particularly wastewater contributions from commercial or industrial sources. The expected range of COD/BOD ratios is between 1.5 and 3.5 (Henze *et al.* 2008).

Inventory compilers should compare country-specific data on COD/BOD ratio in domestic wastewater to IPCC default values of 2.4. If inventory compilers use country-specific COD/BOD ratio, they should provide documented justification why their country-specific values are more appropriate for their national circumstances.

It should be noted that the emission factors are made up of a) the maximum CH₄ producing capacity (B₀) and b) MCF which indicates the extent to which the B₀ is released in each wastewater treatment/discharge pathways or systems. If a country chooses to introduce country-specific data for B₀ based on measured composition of wastewater, they must also update the MCF because the MCFs were developed using the default B₀ values.

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Annex 6A.3 Estimation of default methane conversion factors for CH₄ in centralised wastewater treatment plants treating domestic wastewater

We reviewed scientific literature to find measured CH₄ emissions from full-scale domestic wastewater treatment plants (excluding anaerobic sludge digestion). Although much literature exists, relatively few studies relate to full-scale treatment plants and provide key information such as influent organics in wastewater or organics removed in sludge. We reviewed and accepted data from 14 wastewater treatment plants. MCFs were calculated from data presented in studies, including measured CH₄ emissions from wastewater treatment plant (not including sludge digesters), influent organics in wastewater, estimates of organics removed in sludge, and the IPCC default B₀ value. Some data were not accepted for use because the treatment system co-treated a significant percentage of non-domestic wastewater, such as food waste. The references used, along with the type of treatment system studied and the resulting MCF, are listed in Table 6A.2. Although CH₄ emissions vary by the type of wastewater treatment system used, more on-site exhaustive monitoring data are required to develop different MCFs for different treatment processes. Overall, an MCF of 0.03 was calculated for aerobic wastewater treatment systems with a standard deviation of 0.024 and a range of 0.003–0.09.

TABLE 6A.2 (NEW)		
MCFs BASED ON MEASURED CH ₄ IN FULL-SCALE DOMESTIC WASTEWATER TREATMENT PLANTS		
Type of treatment process (Name of plant)	Reference	MCF ¹
Activated sludge (Stickney)	Bellucci <i>et al.</i> (2010)	0.017
Activated sludge (Northside)	Bellucci <i>et al.</i> (2010)	0.004
Activated sludge (Egan)	Bellucci <i>et al.</i> (2010)	0.014
Activated sludge	Czepiel <i>et al.</i> (1993)	0.013
Biological nutrient removal (Kralingseveer)	Daelman <i>et al.</i> (2013)	0.03
Biological nutrient removal (Kortenoord)	Daelman <i>et al.</i> (2013)	0.02
Biological nutrient removal (Papendrecht)	Daelman <i>et al.</i> (2013)	0.04
Sequencing batch reactor (Holbæk)	Delre <i>et al.</i> (2017)	0.038
Activated sludge (Källby)	Delre <i>et al.</i> (2017)	0.048
Biological nutrient removal (Lundtofte)	Delre <i>et al.</i> (2017)	0.014
Biological nutrient removal (Lynetten)	Delre <i>et al.</i> (2017)	0.015
Activated sludge	Kozak <i>et al.</i> (2009)	0.09
5-stage Bardenpho	Kyung <i>et al.</i> (2015)	0.07
Anaerobic/anoxic/oxic (A2O) process	Wang <i>et al.</i> (2011)	0.003
¹ MCF shown was calculated from data presented in studies, including measured CH ₄ emissions from wastewater treatment plant (not including sludge digesters), influent organics in wastewater, estimates of organics removed in sludge, and the IPCC default B ₀ value.		

Annex 6A.4 Calculation of MCF for methane emissions from sewage discharges

Data were obtained from a recent global review article by Deemer *et al.* (2016) for CO₂ and CH₄ emissions from different environments. At equilibrium with current atmospheric conditions, CO₂ is relatively soluble (over 500 µg/L) compared with CH₄ (about 40 ng/L) at 25 degrees Celsius. Therefore, a reliance on studies that only measure diffusive flux will underestimate the relative emission of CH₄. Deemer *et al.* (2016) raised previous estimates of CH₄ emissions by excluding data from sources that did not include assessment of ebullition. Regarding emissions from all kinds of reservoirs, this exclusion had the effect of increasing the relative CH₄ flux (CH₄-C/(CH₄-C + CO₂-C)) to 27% (first data row in Table 6A.3) which is higher than previously thought. For an estimate of this effect on the other data, it is noted that about 52% of the data listed by Deemer and colleagues included ebullition, and for this about 40–60% of the total ebullitive and diffusive flux was due to ebullition. To illustrate the potential effect of adding ebullition to the *reservoir* studies a scaling factor of $1.48 = 0.52/1 + 0.48/0.5$ is applied for *reservoirs* and *hydroelectric reservoirs* in the table. Deemer *et al.* (2016) used (Bastviken *et al.* 2011) for an estimate of methane emissions from *lakes* and the original data (Table 1 in Bastviken and colleagues) suggests a smaller correction factor for that *lake* data of 1.23. For rivers, Stanley *et al.* (2016) reports a relationship similar to Deemer *et al.*'s observation – that of total flux, 46% is ebullitive among those studies reporting both kinds of fluxes (see Table 1 in Stanley *et al.* 2016). Correcting the average of the larger dataset in Stanley *et al.* (2016) with this figure, suggests a range of riverine yields of 0.4–6%. Note that the riverine CO₂ figure in Deemer *et al.* (2016) of 7,954 mg C/m²/day was based on original 1,800 Tg C/y datum in Raymond *et al.* (2013), which has been reduced using the Lauerwald *et al.* (2015) updated estimate of 659 Tg C/y figure and is thus 2,872 mg C/m²/day.

Considering the data in Table 6A.3, the average yield for reservoirs and lakes is 19% (8–27%) while the average for rivers is much lower at 3.5% (0.4–6%). Data ranges rather than standard deviations are given owing to the potential for overlap between the underlying datasets. These CH₄ yield data are used to estimate the MCF. On the same basis as the B₀ calculation, 1.067 kg of O₂ are required to oxidise one kg of glucose. Considering that 40% of glucose is carbon, this implies the COD of glucose on a carbon basis is 2.67 kg COD/kg C. So 0.19 kg CH₄-C/kg (CH₄-C + CO₂-C) corresponds to 0.071 kg CH₄-C/kg COD. Taking a global perspective, Tranvik *et al.* (2009) state that approximately 1.4 Pg of carbon is emitted from inland waters at CO₂ and a further 0.1 Pg as CH₄, the remainder being sequestered in sediments or lost to the ocean. Thus of 3.0 Pg of carbon inputs, 50% is emitted to the atmosphere. Correcting the previous figure for this yield and converting to kg CH₄ suggests $0.071 [\text{kg CH}_4\text{-C/kg COD}] \times (12.01 + 1.00797 \times 4) / 12.01 [\text{kg CH}_4/\text{kg CH}_4\text{-C}] \times 0.5 = 0.048 \text{ kg CH}_4/\text{kg COD}$. For use in conjunction with B₀, then $\text{MCF} = 0.048/0.25 = 0.19$ (0.08–0.27) for lakes and reservoirs. In the same way, $\text{MCF} = 0.035$ (0.004–0.06) for rivers and streams.

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TABLE 6A.3 (NEW)
SUMMARY OF LITERATURE INVESTIGATING METHANE EMISSIONS FROM WASTEWATER DISCHARGE

Type of Waterbody	CH ₄ flux (mg C/m ² /day)	<i>n</i> (reservoirs, lakes or rivers)	CH ₄ flux corrected for ebullition (mg C/m ² /day)	CO ₂ flux (mg C/m ² /day)	Yield	Approx. revised yield	Reference
Reservoirs	120	75	120	330	27%	27%	Deemer <i>et al.</i> (2016)
Reservoirs	82–96	161	121–142	498	14–16%	20–22%	CH ₄ : Bastviken <i>et al.</i> (2011), St. Louis <i>et al.</i> (2000); CO ₂ : St Louis <i>et al.</i> (2000)
Hydroelectric reservoirs	24–112	85, 104	48–224	386–660	6–15%	8–20%	Barros <i>et al.</i> (2011); Li and Zhang (2014)
Lakes	40	66	80	216	16%	19%	CH ₄ : Bastviken <i>et al.</i> (2011); CO ₂ Raymond <i>et al.</i> (2013)
Rivers	6–98	21, 26	11–183	2872	0.21–3.3%	0.39–6.0%	CH ₄ : Bastviken <i>et al.</i> (2011), Stanley <i>et al.</i> 2016; CO ₂ : Lauerwald <i>et al.</i> (2015)

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Annex 6A.5 Estimation of default emission factors for N₂O in domestic wastewater treatment plants

We reviewed scientific literature to find emission factors from full-scale domestic wastewater treatment plants. Although much literature exists regarding N₂O emission from wastewater treatment processes, relatively few studies relate to full-scale treatment plants and provide key information such as influent nitrogen load. We collected 30 data and found that N₂O emission correlated with influent nitrogen load (Figure 6A.1). Emission factors and references are listed in Table 6A.4 and Table 6A.5. Specific and relatively new/uncommon treatment processes such as membrane bioreactors were intentionally excluded from this list in order to develop emission factors for the most typical and widely used treatment processes globally. Although N₂O emissions vary by the type of nitrogen removal process used, more on-site exhaustive monitoring data are required to develop different N₂O emission factors for different treatment processes.

Figure 6A.1 (New) Correlation between influent total nitrogen (TN) loading and N₂O emissions

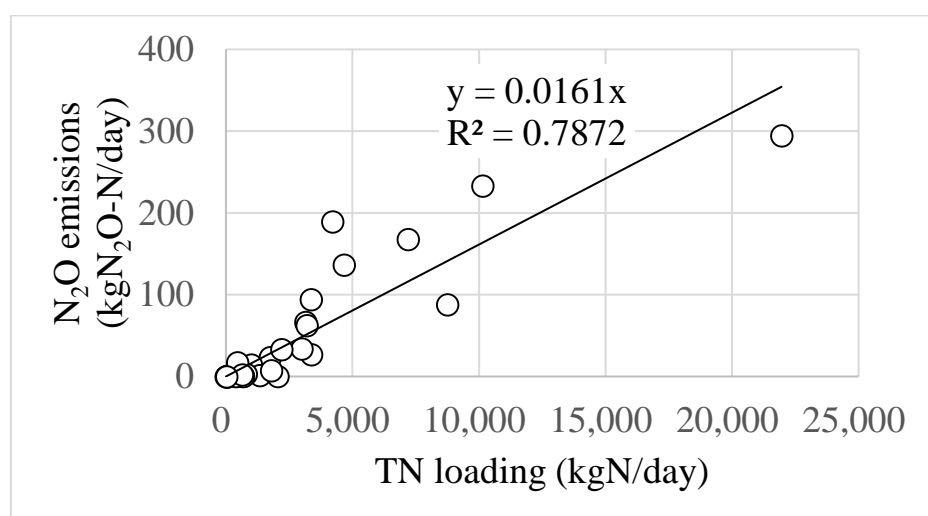


TABLE 6A.4 (NEW)			
DEFAULT N ₂ O EMISSION FACTORS FOR DOMESTIC WASTEWATER TREATMENT PLANTS			
N ₂ O emission factor (kg N ₂ O-N/kg N)	SD	Maximum	Minimum
0.016	0.012	0.045	0.00016

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TABLE 6A.5 (NEW) N₂O EMISSION FACTORS IN FULL-SCALE DOMESTIC WASTEWATER TREATMENT PLANTS			
Type of treatment process¹	Categories	References	N₂O emission factor (kg N₂O-N/kg N)
AO	Biological Nitrogen Removal (BNR)	Daelman <i>et al.</i> (2015)	0.028
AO	BNR	Foley <i>et al.</i> (2010)	0.021
AO	BNR	Foley <i>et al.</i> (2010)	0.045
A2O	BNR	Foley <i>et al.</i> (2010)	0.013
SBR	BNR	Foley <i>et al.</i> (2010)	0.023
OD	BNR	Foley <i>et al.</i> (2010)	0.0080
IA	BNR	Kimochi <i>et al.</i> (1998)	0.0005
EA	BNR	Foley <i>et al.</i> (2010)	0.015
A2O	BNR	Wang <i>et al.</i> (2016)	0.013
CAS	BNR	Aboobakar <i>et al.</i> (2013)	0.00036
AO	BNR	Rodriguez-Caballero <i>et al.</i> (2014)	0.12
OD	BNR	Masuda <i>et al.</i> (2018)	0.00016
AO	BNR	Masuda <i>et al.</i> (2018)	0.0013
AO	BNR	Masuda <i>et al.</i> (2018)	0.0049
Separate-stage BNR	BNR	Ahn <i>et al.</i> (2010)	0.00019
Bardenpho	BNR	Ahn <i>et al.</i> (2010)	0.0036
Step-feed BNR	BNR	Ahn <i>et al.</i> (2010)	0.011
MLE	BNR	Ahn <i>et al.</i> (2010)	0.0007
MLE	BNR	Ahn <i>et al.</i> (2010)	0.0006
OD	BNR	Ahn <i>et al.</i> (2010)	0.0003
Step-feed BNR	BNR	Ahn <i>et al.</i> (2010)	0.015
Step feed, plug flow	BNR	Ni <i>et al.</i> (2015); Pan <i>et al.</i> (2016)	0.019
SBR	BNR	Bao <i>et al.</i> (2016)	0.029
SBR	BNR	Rodriguez-Caballero <i>et al.</i> (2015)	0.038
Plug flow	Non-BNR	Ahn <i>et al.</i> (2010)	0.004
Plug flow	Non-BNR	Ahn <i>et al.</i> (2010)	0.0062
Step-feed non-BNR	Non-BNR	Ahn <i>et al.</i> (2010)	0.0018
Plug flow	Non-BNR	Masuda <i>et al.</i> (2015)	0.023
AO	Non-BNR	Bao <i>et al.</i> (2016)	0.013
IA	Non-BNR	de Mello <i>et al.</i> (2013)	0.0016

¹ AO; Anaerobic-oxic activated sludge process, A2O; Anaerobic-anoxic-oxic activated sludge process, SBR; Sequencing batch reactor, OD; Oxidation ditch, IA; Intermittent aeration process, EA; Extended aeration process, CAS; Conventional activated sludge process, MLE; Modified Ludzack-Ettinger.

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Annex 6A.6 Estimation of default emission factors for N₂O in effluent

Table 6.11 in the 2006 IPCC Guidelines was based on expert judgment. For the refinement of these guidelines, literature was reviewed for reported yields of N₂O from environments with known oxygen saturation conditions. A total of 62 data points for well-oxygenated environments and 59 for low-oxygen environments were obtained from the following literature: Kaplan *et al.* (1978); McElroy *et al.* (1978); Goreau *et al.* (1980); McCarthy *et al.* (1984); Seitzinger *et al.* (1984); Kaplan and Wofsy (1985); Seitzinger (1988); Yoshida (1988); Mantoura *et al.* (1993); Kester *et al.* (1997); Seitzinger and Kroeze (1998); Punshon and Moore (2004); Frame and Casciotti (2010); Beaulieu *et al.* (2011); Yan *et al.* (2012); Zhu *et al.* (2013); Ji *et al.* (2015); Wang *et al.* (2015); He *et al.* (2017).

These data suggested a mean yield of 0.019 kg N₂O-N/kg-N for low oxygen environments (for use in Tier 3 methodologies, if needed) and 0.005 kg N₂O-N/kg-N for well-oxygenated environments (default Tier 1 factor). Calculation of a simple standard deviation is inappropriate for this data, since it is a proportion. Therefore, confidence intervals for these averages were calculated using the adjusted Wald method for small proportions (Bonett & Price 2012). The resulting 95% (two tailed) limits are 0.0041–0.0912 and 0.0–0.0753 respectively.

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Annex 6A.7 List of countries by region included in Table 6.10a

Table 6.10a in Section 6.3.1.3 presents default factors for protein consumed as a fraction of protein supplied, fraction of food not consumed and disposed in sewer, and a fraction to represent additional nitrogen introduced to the sewer from household chemicals. The information is provided by geographic region. The list of countries by region can be found in Table 6A.6.

TABLE 6A.6 (NEW) LIST OF COUNTRIES BY REGION INCLUDED IN TABLE 6.10A	
Region	Country
Europe	Albania, Armenia, Austria, Azerbaijan, Belarus, Belgium, Bosnia and Herzegovina, Bulgaria, Croatia, Cyprus, Czech Republic, Denmark, Estonia, Finland, France, Georgia, Germany, Greece, Hungary, Iceland, Ireland, Italy, Latvia, Lithuania, Luxemburg, Macedonia, Moldova, Montenegro, Netherlands, Norway, Poland, Portugal, Romania, Russian Federation, Serbia, Slovakia, Slovenia, Spain, Sweden, Switzerland, Ukraine, United Kingdom
North America and Oceania	Australia, Canada, New Zealand, United States of America
Industrialized Asia	Japan, China, South Korea
Sub-Saharan Africa	Angola, Benin, Botswana, Burkina Faso, Burundi, Cameroon, Central African Rep., Chad, Congo-Brazzaville, Congo-Kinshasa, Cote d'Ivoire, Equatorial Guinea, Eritrea, Ethiopia, Gabon, Gambia, Ghana, Guinea, Guinea-Bissau, Kenya, Lesotho, Liberia, Malawi, Mali, Mauritania, Mozambique, Namibia, Niger, Nigeria, Rwanda, Senegal, Sierra Leone, Somalia, South Africa, Sudan, Swaziland, Tanzania, Togo, Uganda, Zambia, Zimbabwe
North Africa, West and Central Asia	Algeria, Egypt, Iraq, Israel, Jordan, Kazakhstan, Kuwait, Kyrgyzstan, Lebanon, Libya, Mongolia, Morocco, Oman, Saudi Arabia, Syria, Tajikistan, Tunisia, Turkey, Turkmenistan, Utd Arab Emirates, Uzbekistan, Yemen
South and Southeast Asia	Afghanistan, Bangladesh, Bhutan, Cambodia, India, Indonesia, Iran, Laos, Malaysia, Myanmar, Nepal, Pakistan, Philippines, Sri Lanka, Thailand, Vietnam
Latin America	Argentina, Belize, Bolivia, Brazil, Chile, Colombia, Costa Rica, Cuba, Dominican Rep., Ecuador, El Salvador, Guatemala, Guyana, Haiti, Honduras, Jamaica, Mexico, Nicaragua, Panama, Paraguay, Peru, Suriname, Uruguay, Venezuela

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