

CHAPTER 3

CROSS-CUTTING GUIDANCE ON REWETTED PEATLANDS AND ORGANIC SOILS

Contents

3	Cross-cutting guidance on REWETTED PEATLANDS AND organic soils.....	3
3.1	Introduction	3
3.2	Greenhouse Gas Emissions and Removals from Re-wetted Peatlands and Organic Soils	4
3.2.1	CO ₂ Emissions and Removals from Re-wetted Peatlands and Organic Soils	4
3.2.2	CH ₄ Emissions and Removals from Re-wetted Peatlands and Organic Soils	10
3.2.3	N ₂ O Emissions and Removals from Re-wetted Peatlands and Organic Soils	12
3.3	Completeness, Time series, consistency, and QA/QC.....	13
3.3.1	Completeness	13
3.3.2	Developing a consistent time series	13
3.3.3	Quality Assurance and Quality Control (QA/QC)	14
3.3.4	Reporting and Documentation	14
3.4	Basis for future methodological development	Error! Bookmark not defined.
Annex 3A.1	Estimation of default emission factors for CO ₂ -C in rewetted peatlands and organic soils.....	18
Annex 3A.2	Estimation of default emission factors for CO ₂ -DOC in rewetted peatlands and organic soils..	19
Annex 3A.3	Estimation of default emission factors for CH ₄ -C in rewetted peatlands and organic soils.....	20

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26

Equations

27	Equation 3.1 CO ₂ -C EMISSIONS/REMOVALS FROM REWETTED PEATLANDS AND ORGANIC SOILS	4
28	Equation 3.2	5
29	Equation 3.3	8
30	Equation 3.4	10
31	Equation 3.5	10
32	Equation 3.6	13
33		

34

Figures

35	Figure 3.1 Decision tree to estimate CO ₂ -C emissions/removals from rewetted peatlands and organic soils.	
36	7	
37	Figure 3A.1	18

38

Tables

39	Table 3.1	4
40	Table 3.2 Emission Factors (EF _{CO₂-C rewetted} , in tonnes CO ₂ -C ha ⁻¹ yr ⁻¹) for CO ₂ -C (± Standard Deviation) for	
41	Rewetted Peatlands and Organic Soils	8
42	Table 3.3 Default DOC emission factors (EF _{DOC rewetted} in tonnes CO ₂ -C ha ⁻¹ yr ⁻¹) for re-wetted peatlands	
43	and organic soils (values in parentheses represent uncertainty ranges).....	9
44	Table 3.4 Default emission factors for CH ₄ from rewetted peatlands and organic soils (all values in kg CH ₄ -	
45	C ha ⁻¹ yr ⁻¹).....	12
46	Table 3A.1 DOC concentration (above) or flux (below) comparisons between drained and re-wetted peats	
47	19
48		
49		

50 **3 CROSS-CUTTING GUIDANCE ON REWETTED** 51 **PEATLANDS AND ORGANIC SOILS**

52 **3.1 INTRODUCTION**

53 **What is rewetting, restoration, rehabilitation and how they affect GHG**

54 For the purpose of this chapter, re-wetting, restoration and rehabilitation are understood as follows:

55 Wetlands are lands characterised by water saturation of the soil dominating hydrological and biogeochemical
56 processes. Rewetting is the action of raising the water table on drained land to re-establish such conditions, e.g.
57 by blocking drainage ditches or disabling pumping facilities.

58 Restoration is the permanent re-establishment, on formerly drained sites, of hydrological and biogeochemical
59 processes characteristics of saturated soils, as well as of the vegetation cover that pre-dated the drainage of these
60 soils. Restoration necessarily includes the re-wetting of formerly drained areas.

61 Rehabilitation or reclamation is the re-establishment, on formerly drained sites, of some of – but not necessarily
62 all - the hydrological, biogeochemical and ecological processes and functions that characterized pre-drainage
63 conditions. As such, rehabilitation can involve a large variety of practices on formerly drained peatlands or
64 organic soils, which may or may not include re-wetting.

65 **Scope of this guidance: wetland types covered, gases, pools, relevant classifiers**

66 The guidance provided in this chapter will include re-wetting and restoration of peatlands and organic soils,
67 excluding other forms of rehabilitation or reclamation, whose outcome is more variable and site-specific. This
68 chapter seeks to avoid repeating guidance already provided, hence wherever appropriate will refer to already
69 existing guidance in these Guidelines. Table 3.1 summarizes the coverage of this chapter.

70 In keeping with the recommendations in chapter 7, volume 4, 2006 IPCC Guidelines, GHG emissions from re-
71 wetted and restored peatlands or organic soils are quantified as fluxes rather than C stock changes. Likewise,
72 carbon emissions or removals in the form of CO₂ will be described as CO₂-C, and when they occur in the form
73 of CH₄ as CH₄-C.

74 Contrary to most ecosystems, the distinction between C pools in some wetlands can be difficult, especially
75 between the dead organic matter and soil pools. For example, the dead portion of moss species characteristics of
76 many northern peatlands could be included in the dead organic matter or soil pool. Moreover, the default
77 emission factors in this chapter were all derived from flux measurements; these fluxes, when measured over
78 moss-covered or grass-covered peatlands or organic soils, integrate all C pools. In all cases the guidance in this
79 chapter will clarify which C pools are included in default EFs. Peatlands and organic soils can also support
80 perennial woody vegetation; the default approach will treat this carbon pool as in the remainder of the 2006 GLs.

81 Evidence of successfully re-wetted peatlands or organic soils in tropical or sub-tropical regions was insufficient
82 to provide a sound basis for the development of default EFs. Specifically, no evidence could be found of success
83 in raising the water table to the level observed in pristine peatlands. As a result, flux measurements from pristine
84 peatlands in tropical or sub-tropical climates could be used as a proxy for EF on re-wetted peatlands in these
85 regions and only provided for CO₂ as a basis for future methodological development. The assumption remains
86 that GHG emissions will respond to re-wetting; however countries in tropical and sub-tropical regions where
87 significant areas of peatlands or organic soils have been re-wetted should be encouraged to develop science-
88 based, documented, country-specific emission factors for both CO₂ and CH₄ emissions.

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TABLE 3.1

Climate region \ Peat type	Bog	Fen	Tropical peat
Boreal/temperate			
CO ₂	Section 3.2.1	Section 3.2.1	-
CH ₄	Section 3.2.2	Section 3.2.2	-
N ₂ O	Section 3.2.3	Section 3.2.3	-
Tropical			
CO ₂	-	-	Appendix
CH ₄	-	-	Section 3.4.1
N ₂ O	-	-	Section 3.4.2

89

90 How to use guidance in this chapter and relationship to reporting categories

91 The post-rewetting land use of peatlands and organic soils can vary according to national circumstances.
 92 Depending on the management practices and the national land-use classification, it may be more appropriate to
 93 report the GHG estimates of re-wetted peatlands or organic soils under forest land, cropland, grassland or
 94 wetlands. It is therefore recommended to consider this guidance as potentially common to several reporting
 95 categories. Because the functioning of these ecosystems have already been deeply altered, reporting re-wetted
 96 peatlands or organic soils as unmanaged land is not consistent with good practice.

97 3.2 GREENHOUSE GAS EMISSIONS AND 98 REMOVALS FROM RE-WETTED PEATLANDS 99 AND ORGANIC SOILS

100 3.2.1 CO₂ Emissions and Removals from Re-wetted 101 Peatlands and Organic Soils

102 CO₂-C emissions/removals from re-wetted peatlands and organic soils have the following components:

EQUATION 3.1
CO₂-C EMISSIONS/REMOVALS FROM REWETTED PEATLANDS AND ORGANIC SOILS

$$\text{CO}_2\text{-C}_{\text{rewetted}} = \text{CO}_2\text{-C}_{\text{soil}} + \text{CO}_2\text{-C}_{\text{veg}} + \text{CO}_2\text{-C}_{\text{woody biomass}} + \text{CO}_2\text{-C}_{\text{DOM}} + \text{CO}_2\text{-C}_{\text{DOC}}$$

107 Where:

108 $\text{CO}_2\text{-C}_{\text{rewetted}}$ = CO₂-C emissions/removals from re-wetted peatlands and organic soils, tonnes C yr⁻¹

109 $\text{CO}_2\text{-C}_{\text{soil}}$ = CO₂-C emissions from the soil, tonnes C yr⁻¹

110 $\text{CO}_2\text{-C}_{\text{veg}}$ = net CO₂-C emissions by, or removals from re-wetted vegetation, tonnes C yr⁻¹

111 $\text{CO}_2\text{-C}_{\text{woody biomass}}$ = net CO₂-C emissions by, or removals from perennial woody biomass, tonnes C yr⁻¹

112 $\text{CO}_2\text{-C}_{\text{DOM}}$ = CO₂-C emissions from dead organic matter, tonnes C yr⁻¹

113 $\text{CO}_2\text{-C}_{\text{DOC}}$ = CO₂-C emissions from dissolved organic carbon, tonnes C yr⁻¹

114 $\text{CO}_2\text{-C}_{\text{soil}}$ is produced during the decomposition of the soil/peat by heterotrophic organisms and is controlled
 115 strongly by oxygen availability within the soil. $\text{CO}_2\text{-C}_{\text{veg}}$ is the net result of two processes by the vegetation
 116 component: photosynthesis (CO₂ uptake) and autotrophic respiration (CO₂ emissions).

117 $\text{CO}_2\text{-C}_{\text{DOM}}$ is produced from the decomposition of dead organic matter, such as litter, root exudates and coarse
 118 woody debris.

119 The default EFs provided in this section capture the components $\text{CO}_2\text{-C}_{\text{soil}}$, $\text{CO}_2\text{-C}_{\text{veg}}$ and $\text{CO}_2\text{-C}_{\text{DOM}}$ together
 120 rather than C stock changes by pool. These default EFs are developed from annualized flux measurements,
 121 ensuring that there is no seasonal bias.

122

123 This chapter does not provide default methodologies for estimating C stock changes in perennial woody biomass
 124 on re-wetted peatlands or organic soils. When the vegetation on re-wetted sites consists of perennial woody
 125 biomass, guidance in vol 4, chapter 2, 4, 5 and 6 should be used.

126

127 $\text{CO}_2\text{-C}_{\text{DOC}}$ is produced from the decomposition of dissolved organic carbon (DOC) lost via aquatic pathways.
 128 DOC is the major component of waterborne carbon export, Particulate Organic Carbon (POC), dissolved CO_2
 129 and Dissolved Inorganic Carbon (DIC) being the other minor components (Billett *et al.* 2004, Billett and Moore
 130 2008, Worrall *et al.* 2009, Billett *et al.* 2010, Dinsmore *et al.* 2010). Re-wetting especially of non-re-vegetated
 131 cutaway peatlands could lead to high POC export which can be converted to CO_2 . However, POC is likely to be
 132 simply translocated from the re-wetted peatlands to other stable carbon stores, such as freshwater or marine
 133 sediments and may not lead to CO_2 emission. Due to the uncertain fate of POC export, no estimation
 134 methodology is presented in this text. Finally, dissolved CO_2 or DIC emissions form a relatively small
 135 component of total land-atmosphere CO_2 exchange.

136

137 CHOICE OF METHOD

138 Tier 1

139 The decision tree in Figure 3.1 presents guidance in the selection of the appropriate tier for the estimation of
 140 greenhouse gas emissions from re-wetted peatlands and organic soils. Under Tier 1, the nationally derived area
 141 of rewetted peatlands and organic soils is multiplied by an emission factor, which is disaggregated by climate
 142 region and peatland type (Equation 3.2).

143

EQUATION 3.2

144

$$145 \text{CO}_2\text{-C}_{\text{rewetted}} = (A_{\text{rewetted}} \cdot EF_{\text{CO}_2 \text{ rewetted}}) + (A_{\text{rewetted}} \cdot EF_{\text{DOC rewetted}})$$

146 Where:

147 $\text{CO}_2\text{-C}_{\text{rewetted}}$ = $\text{CO}_2\text{-C}$ emissions/removals from rewetted peatlands and organic soils, tonnes C yr^{-1})

148 A_{rewetted} = total area of peatlands and organic soils that have been rewetted, ha

149 $EF_{\text{CO}_2 \text{ rewetted}}$ = $\text{CO}_2\text{-C}$ emission factor for rewetted peatlands and organic soils, tonnes C $\text{ha}^{-1} \text{yr}^{-1}$

150 $EF_{\text{DOC-rewetted}}$ = $\text{CO}_2\text{-C}$ emission factor for DOC from a re-wetted peatlands and organic soils tonnes C ha^{-1}
 151 yr^{-1}

152 Tier 2

153 Tier 2 calculations use country-specific emission factors and parameters, spatially disaggregated to reflect
 154 regionally important practices and dominant ecological dynamics. It may be appropriate to sub-divide activity
 155 data and emission factors according to the present vegetation composition (Couwenberg *et al.* 2011) and
 156 previous land use history. $\text{CO}_2\text{-C}$ fluxes during the time period immediately following rewetting (0-5 years) have
 157 been shown to be significantly different to those observed 5 years or more after rewetting (Annex or in Guidance
 158 for future methodology to be provided). Countries are encouraged to develop more detailed EFs for rewetted
 159 peatlands and organic soils by the use of a 5 year “transition period”.

160 A Tier 2 approach to quantify DOC fluxes may follow the Tier 1 methodology provided above, but should utilise
 161 country-specific information where possible to refine the emission factor used as well as the conversion factor.
 162 Refinements could include:

- 163 (i) Use of country-level measurements from re-wetted peatlands with various restoration
 164 techniques and initial status. Direct measurements of DOC fluxes from re-wetted peats could replace
 165 the $\text{DOC}_{\text{FLUX_NATURAL}}$ values used in the Tier 1 default approach, i.e. replacing the default assumption
 166 that re-wetted peatlands revert to pre-drainage DOC fluxes. Differences between pre-drainage and re-
 167 wetted DOC fluxes could occur due to the presence or absence of vegetation on re-wetted sites; if re-
 168 wetted vegetation composition differs from natural peatlands, or due to factors associated with peat
 169 restoration techniques, such as the creation of pools behind dams, or the application of mulch to
 170 support vegetation re-establishment, which has been shown to have a direct impact on DOC fluxes

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- 171 (Strack *et al.* 2011), or the use of biomass to infill ditches. The use of an initial EF (between 0 and 5
172 years after restoration) and 5 years or more after rewetting (as per CO₂ emissions) is likely to be
173 applied given the dynamics of the vegetation colonisation/development.
- 174 (ii) Use of country-level measurements from natural peatlands to obtain accurate values of
175 DOC_{FLUX_NATURAL} for that country, for example by developing specific values for raised bogs versus
176 fens, or for blanket bogs under different precipitation levels (see equation used at the bottom of Table
177 3.2).
- 178 (iii) Use of alternative values for F_{DOC-CO2} where evidence is available to indicate that a different
179 proportion of DOC exported from re-wetted peatlands and organic soils is transferred to stable long-
180 term carbon stores, such as lake or marine sediments.

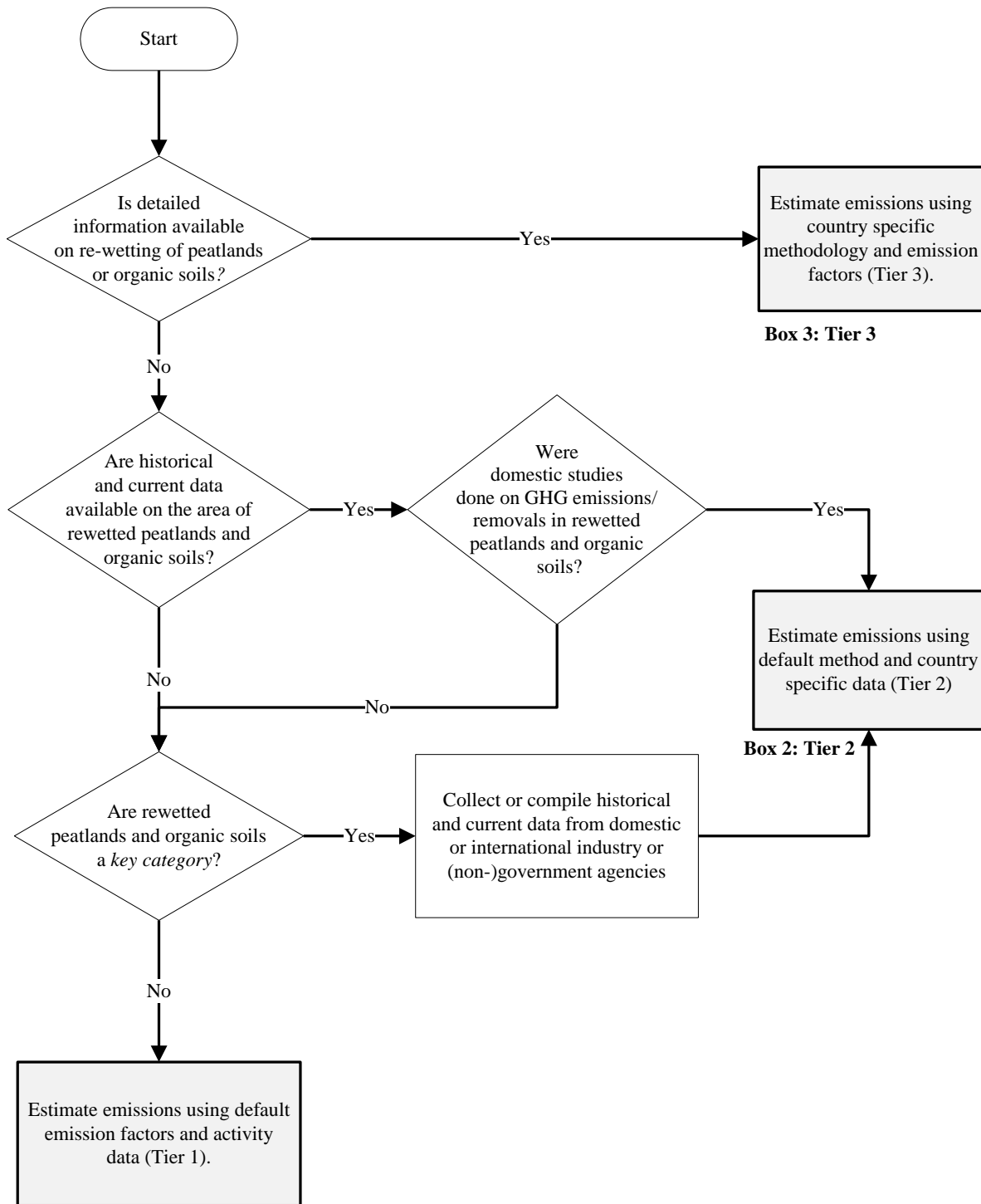
181 **Tier 3**

182 A Tier 3 approach involves a comprehensive understanding and representation of the dynamics of CO₂-C
183 emissions and removals on re-wetted peatlands and organic soils, including the effect of site characteristics, peat
184 type and depth, vegetation composition, soil temperature and mean water table depth (WTD). These parameters
185 could also be used to describe DOC release using process-based models. A Tier 3 approach might also include
186 the entire DOC export from re-wetted sites and consideration of the temporal variability in DOC release in the
187 years following re-wetting (e.g. initial period of less than 5 years) which will be also dependent on the
188 restoration techniques used.

189

190
191

Figure 3.1 Decision tree to estimate CO₂-C emissions/removals from rewetted peatlands and organic soils.



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194 **CHOICE OF EMISSION FACTORS**195 **Tier 1**

196 The implementation of the Tier 1 method requires the application of default emission factors EF_s provided in
 197 Table 3.2, where they are disaggregated by climate region (boreal, temperate) and peatland type (bog, fen).
 198 There is insufficient data available to permit the inclusion of rewetted tropical peatlands at this time. Peatland
 199 type can be inferred from the nutrient status of the peat substrate: e.g. *sphagnum* peat is characteristic of the
 200 more nutrient poor (oligotrophic) bogs and *Phragmites* peat is common in nutrient rich (minerotrophic) fens.
 201 Countries with significant areas of re-wetted peatlands and organic soils are encouraged to develop domestic
 202 emission factors and develop estimates of emissions and/or removals using Tier 2 or 3 methodologies.

203

204 The derivation of these default emissions factors values is fully described in Annex 3.1. In summary, while
 205 available data indicated that CO₂ fluxes from natural peatlands and re-wetted peatlands are in both cases
 206 correlated with water table, it was ascertained that this relationship was not significantly different between the
 207 two ecosystems (except for temperate fens). Therefore CO₂ fluxes from natural undrained sites were used in
 208 addition to CO₂ fluxes from re-wetted sites in order to provide robust estimation of EF_s as shown in Table 3.3. In
 209 the case of temperate fens, available data from re-wetted sites only were used to estimate the EF.

210

	BOREAL	TEMPERATE
BOGS (raised and blanket)	-0.26 (±0.48)	-0.52 (±1.17)
FENS	-0.93 (±0.94)	+0.72 (±3.42)

211

212 **EF_{DOC-rewetted}**

213 Robust data show that natural, undrained peatlands export some DOC and these fluxes increase following
 214 drainage (see Chapter 2). Available data from re-wetted sites suggest that the level of DOC reduction after re-
 215 wetting equates to the DOC increase after drainage. Consequently, it is assumed that re-wetting leads to a
 216 reversion to natural DOC flux levels (see Annex 3.2 for explanation). Therefore, to make best use of available
 217 data, EF_s for re-wetted peatlands have been calculated according to Equation 3.3:

218

219

$$\text{EF}_{\text{DOC re-wetted}} = \text{DOC}_{\text{FLUX NATURAL}} * F_{\text{DOC-CO}_2}$$

220 Where:

221

EF_{DOC-Re-wetted} = emission factor for DOC from a re-wetted site, tonnes C m⁻² yr⁻¹

222

DOC_{FLUX NATURAL} = Flux of DOC from natural (undrained) peatland, tonnes C m⁻² yr⁻¹

223

F_{DOC-CO₂} = Conversion factor for proportion of DOC converted to CO₂ following export from site

224 EF_{DOC-re-wetted} values (required in Equation 3.2) are provided in Table 3.3. The DOC_{FLUX NATURAL} values were
 225 derived based on available data indicating that the rate of DOC export from temperate and boreal raised bogs and
 226 fens is positively correlated with water fluxes. A simple schema for deriving values of DOC_{FLUX NATURAL} as a
 227 function of rainfall was used and provides a separate EF as shown in Table 3.3. This relationship was not
 228 ascertained for blanket bogs, which display higher water fluxes in general and, therefore, a separate category
 229 using all DOC flux values from natural blanket bogs in the literature has been used and associated EF computed.
 230 This was also done separately for tropical peatlands. The derivation of these values is fully described in Annex X.

231

232 An understanding of the ultimate fate of DOC export, i.e. whether it is returned to the atmosphere as CO₂ (or
 233 even CH₄), is still poor and yet of great significance in terms of GHG reporting. The parameter F_{DOC-CO₂} sets the
 234 proportion of DOC exported from peatlands and organic soils which is ultimately converted to CO₂. A value of
 235 zero would coincide with all the DOC export being deposited in stable forms in lake or marine sediments, as this
 236 would simply represent a translocation of carbon between stable stores, it would not need to be estimated.
 237 However, most data on DOC processing do indicate that a high proportion is converted to CO₂ in headwaters,

238 rivers, lakes and coastal seas (see Chapter 2). Therefore, a Tier 1 default $F_{\text{DOC-CO}_2}$ value of 90% is proposed, with
 239 an uncertainty range of 80-100%.

Peat type	Precipitation regime (mm yr^{-1})	$\text{DOC}_{\text{FLUX_NATURAL}}$ (tonnes $\text{C m}^{-2}\text{ yr}^{-1}$)	$EF_{\text{DOC_rewetted}}$
Temperate/boreal raised bog/fen^a	< 500	5	5 (3-8)
	500-700	12	10 (7-15)
	700-900	18	16 (12-21)
	> 900	24	22 (17-36)
Temperate blanket bog	All	21	19 (10-28)
Tropical	All	59	53 (38-69)

a $\text{DOC}_{\text{FLUX_NATURAL}}$ values for boreal/temperate raised bogs and fens are calculated from the equation $\text{DOC}_{\text{FLUX_NATURAL}} = (0.0317 \cdot \text{Precipitation}) - 7.5$, in the units shown above, using representative mid-range values of 400, 600, 800 and 1000 mm for each of the classes shown. This equation may be used to calculate site-specific estimates where actual precipitation values are known.

240 **Tier 2 and 3**

241 The uncertainty of EFs can be reduced on the basis of spatially disaggregated CO_2 flux measurements. Flux
 242 measurements (eddy covariance or chamber methods) should be carried out at temporal resolutions sufficiently
 243 defined to capture as wide a range as possible of the abiotic (e.g. irradiation, soil temperature, WTD) and biotic
 244 (e.g. vegetation composition) factors that drive CO_2 dynamics in re-wetted peatlands and organic soils.

245

246 Countries where perennial woody biomass plays a significant role in the net $\text{CO}_2\text{-C}$ exchange between re-wetted
 247 peatlands or organic soils and the atmosphere should develop country-specific factors reflecting C stock changes
 248 in this pool under typical management practices. Guidance to this effect can be found in Vol 4, chapters 3-5.

249

250 **CHOICE OF ACTIVITY DATA**

251 All Tiers require data on areas of re-wetted peatlands or organic soils, broken down by climate zone, type of
 252 peatland or organic soils.

253

254 **Tier 1**

255 The default methodology assumes that a country has data on the area of re-wetted peatlands or organic soils and
 256 the type of peatland or organic soils, consistent with the advice above on the selection of emission factors. Data
 257 can be obtained from the peat extraction industry, forestry or agricultural agencies, as well as from government
 258 and non-government sources. Re-wetting plans and land use maps are useful sources of information.

259 **Tier 2 and 3**

260 Tier 2 and 3 approaches are likely to involve a more detailed stratification than in Tier 1. This can include further
 261 sub-divisions based on time since re-wetting, previous land use history and current land use as well as vegetation
 262 composition. It is good practice to further sub-divide default classes based on empirical data that demonstrates
 263 significant differences in $\text{CO}_2\text{-C}$ fluxes among the proposed categories. For application of a direct measurement-
 264 based inventory in Tier 3, similar or more detailed data on the combinations of climate, soil, topographic and
 265 management data are needed, relative to the Tiers 1 and 2 methods, but the exact requirements will depend on
 266 the measurement design and country-specific circumstances.

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3.2.2 CH₄ Emissions and Removals from Re-wetted Peatlands and Organic Soils

CH₄ emissions and removals from the soils of re-wetted peatlands and organic soils result from 1) emissions or removals resulting from the balance between CH₄ production and oxidation in the soil column and 2) emission of CH₄ produced by the combustion of organic matter during fire (Equation 3.4).

EQUATION 3.4

$$CH_4-C_{rewetted} = CH_4-C_{soil} + CH_4-C_{biomass\ burn} + CH_4-C_{soil\ burn}$$

Where:

CH₄-C_{rewetted} = CH₄-C emissions from re-wetted peatlands or organic, tonnes C yr⁻¹

CH₄-C_{soil} = emissions of CH₄-C from peatlands or organic soils subject to rewetting, tonnes C yr⁻¹

CH₄-C_{biomass burn} = emissions of CH₄-C from biomass burning on rewetted peatlands or organic soils, tonnes C yr⁻¹

CH₄-C_{soil burn} = emissions of CH₄-C from soil or peat burning on rewetted peatlands or organic soils, tonnes C yr⁻¹

The default EFs provided in this section will only cover CH₄-C_{soil}. Generic methodologies for estimating CH₄ emissions from the burning of vegetation, dead organic matter (CH₄-C_{biomass burn}) are provided in chapter 2 of volume 4, while methodologies specific to vegetation and DOM burning in forest land, cropland, grassland and wetlands are provided in chapters 3-6 of volume 4. When burning occurs, emissions from the burning of soil should be estimated using country-specific (Tier 2) emission factors.

As noted in Chapter 2, emissions of CH₄-C from drainage ditches can be much higher than the surrounding drained peat fields. Few data are available on CH₄-C emissions from ditches of rewetted peatlands and organic soils and in some cases ditches may be filled during rewetting activities. When ditches remain, countries are encouraged to include estimates of CH₄-C ditch emissions using methodology provided in Chapter 2 (Equation 2.X) and country specific emission factors. Table 2.X can also be consulted for guidance on emission factors for ditches in drained peatlands.

As noted in Chapter 2, emissions of CH₄-C from drainage ditches can be much higher than the surrounding drained peat fields. Few data are available on CH₄-C emissions from ditches of rewetted peatlands and organic soils and in some cases ditches may be filled during rewetting activities. When ditches remain, countries are encouraged to include estimates of CH₄-C ditch emissions using methodology provided in Chapter 2 (Equation 2.X) and country specific emission factors. Table 2.X can also be consulted for guidance on emission factors for ditches in drained peatlands.

CHOICE OF METHOD

Refer to Figure 3.1 for the decision tree to select the appropriate tier for the estimation of estimate CH₄ emissions or removals from rewetted peatland or organic soils.

Tier 1

The default methodology covers CH₄ emissions and removals from rewetted peatlands soils and re-wetted organic soils (Equation 3.5).

EQUATION 3.5

$$CH_4-C_{soil} = \left[\frac{(A_{rewetted} * EF_{CH_4\ soil})}{1000} \right]$$

Where:

CH₄-C_{soil} = methane emissions from the soils of re-wetted peatlands and organic soils, tonnes C ha⁻¹ yr⁻¹

A_{rewetted} = total area of peatlands and organic soils that have been rewetted, ha yr⁻¹

EF_{CH₄ soil} = emission factor from rewetted peatlands or organic soil, kg CH₄-C ha⁻¹ yr⁻¹

312 Rewetted area should be subdivided by climate zone (boreal or temperate) and appropriate emission factors
313 applied. The data available at present is insufficient to support the development of default EF for tropical regions.
314 When information is available on the peatland type, countries are encouraged to further subdivide rewetted area
315 into bogs and fens, multiply each one by the appropriate emission factor and sum the products for the total CH₄
316 emission or removal.

317 **Tier 2**

318 Tier 2 calculations use country-specific emission factors and parameters, spatially disaggregated to reflect
319 regionally important practices and dominant ecological dynamics. In general, methane fluxes from wet
320 peatlands are extremely skewed, approaching a log-normal (right-tailed) distribution. This asymmetry towards
321 rare, but high efflux values causes high mean values compared to the most likely encountered median values.
322 Countries are encouraged to further scientific understanding of methane effluxes from re-wetted peatland sites,
323 allowing for more precise estimation.

324

325 Methane fluxes from peat and organic soils strongly depend on the depth of the water table, with efflux
326 increasing steeply from near zero when mean annual water levels stands below -20 cm from the surface, to very
327 variable and high values at water levels less than 20 cm from the surface. Variability is even greater on flooded
328 sites, where both low (Couwenberg & Fritz 2012; Couwenberg et al., 2010) and high (Augustin & Chojnicki
329 2008; Glatzel et al., 2011) flux values have been observed.

330

331 Prior land-use (e.g. agriculture, peat extraction, forestry) can influence CH₄ fluxes from re-wetted peatlands or
332 organic soils. For example, CH₄ emissions following the re-wetting of some agricultural land with nutrient
333 enriched top-soil appear higher compared to average emission factors (Augustin & Chojnicki, 2008; Hahn-
334 Schöfl et al., 2011; Glatzel et al., 2011) whereas rewetted boreal cutover peatlands may have CH₄ emissions
335 below the average emission factors (Waddington and Day, 2007). It may therefore also be appropriate to
336 subdivide activity data and emission factors according to previous land-use when this information is available.
337 The influence of previous land use may diminish over time.

338

339 Data on CH₄ emissions and removals from rewetted peatlands and organic soils remains relatively sparse,
340 particularly at locations rewetted for longer than 10 years. Due to the limited number of long-term rewetting
341 studies, the change in CH₄ emissions over time remains unclear. The pattern of changes in CH₄ flux with time
342 since rewetting may be linked to prior land-use. Research on restored cutover peatlands in Canada indicates an
343 increase in CH₄ emissions each year during the first three years post-restoration as the emerging vegetation
344 cover provides fresh substrate for CH₄ production (Waddington and Day, 2007). In contrast, rewetting of
345 intensive grassland on fen peat suggests that CH₄ emissions may decline over time as litter inundated during
346 rewetting activities is rapidly decomposed in the first few years (Augustin and Joosten 2007). Changes in CH₄
347 emissions and removals over time are likely linked to vegetation succession (e.g. Tuittila et al. 2000) and thus
348 understanding the pattern of emissions over time will likely require the inclusion of vegetation information.

349

350 Several studies in both undisturbed and rewetted peatlands indicate the important role that vegetation may play
351 for providing substrate for CH₄ production and for transporting CH₄ from the saturated soil to the atmosphere
352 (e.g. Bubier 1995; Shannon et al. 1996; Marnier et al. 2004; Tuittila et al. 2000; Wilson et al. 2007; Dias et al.
353 2010). Species known to transport CH₄ from the soil to the atmosphere include, but are not limited to *Calla*,
354 *Carex*, *Cladium*, *Eleocharis*, *Equistem*, *Eriophorum*, *Glyceria*, *Nuphar*, *Nymphaea*, *Peltandra*, *Phalaris*,
355 *Phragmites*, *Sagittaria*, *Scheuchzeria*, *Scirpus*, and *Typha* (Sebacher et al. 1985, Chanton et al. 1992, Schimel
356 1995, Shannon et al. 1996, Frenzel & Rudolph 1998, Verville et al. 1998, Yavitt & Knapp 1998, Grünfeld &
357 Brix 1999, Frenzel & Karofeld 2000, Tuittila et al. 2000, Arkebauer et al. 2001, Armstrong & Armstrong 2011,
358 Askaer et al. 2011). The presence of these aerenchymous shunt species has a marked effect on methane efflux
359 from peatlands (Couwenberg & Fritz, 2012). Most of the data available at present are from sites that were
360 rewetted only recently when methane fluxes were measured. Countries are encouraged to use methane flux
361 measurements from longer term rewetted sites and develop nationally specific emission factors that address
362 vegetation composition (see Couwenberg et al., 2011).

363

364 **Tier 3**

365 A Tier 3 approach involves a comprehensive understanding and representation of the dynamics of CH₄
366 emissions on re-wetted peatlands and organic soils, including the representation of interactions between the
367 dominant drivers of CH₄ dynamics, as described above.

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368

CHOICE OF EMISSION FACTORS**Tier 1**

The implementation of the Tier 1 method requires the application of default emission factors $EF_{CH_4 \text{ soil}}$ provided in Table 3.4, where they are disaggregated by climate region (boreal, temperate) and peatland type (bog, fen). Available flux data for rewetted tropical peatlands are insufficient for the development of default EFs. Comparison of hourly flux data suggests that efflux in SE Asian peat swamps (the largest extent of tropical peatland) will be much lower than in temperate/boreal regions (Couwenberg et al., 2010). Annex 3.3 provides more details on the derivation of these default EFs.

Climate zone	EF _{rewetted}	Uncertainty	Peatland type	EF _{rewetted}	Uncertainty
Boreal	80 (n= 27)		Bog	16 (n=12)	
			Fen	162 (n=12)	
Temperate	265 (n=104)		Bog	156 (n=41)	
			Fen	374 (n=63)	
Tropical	p.m.				

Tier 2 and 3

The uncertainty of emission factors can be reduced by using country specific emission factors for each climate zone and peatland type. It can be further reduced by implementing scientific findings relating CH₄ emissions to prior land use, the depth of the water table, vegetation cover and time since re-wetting.

381

CHOICE OF ACTIVITY DATA

All Tiers require data on areas of re-wetted peatlands or organic soils, broken down by climate zone, type of peatland or organic soils.

Tier 1

The default methodology assumes that a country has data on the area of re-wetted peatlands or organic soils and the type of peatland or organic soils, consistent with the advice above on the selection of emission factors.

388

Tier 2 and 3

More sophisticated estimation methodologies will require the determination of monthly averages of the water table depth; land use and management practices prior to re-wetting; and vegetation composition and the successional changes in vegetation community composition and biomass with time since rewetting. This type of information can be obtained by long-term monitoring of re-wetted sites under various conditions, and should be combined with an enhanced understanding of the processes linking CH₄ emissions to these factors.

395

3.2.3 N₂O Emissions and Removals from Re-wetted Peatlands and Organic Soils

The emissions of N₂O from re-wetted peatlands or organic soils are controlled by the quantity of N available for nitrification and de-nitrification, and the availability of the oxygen required for these chemical reactions. Oxygen availability is in turned controlled by the depth of the water table. Raising the depth of the water table will cause N₂O emissions to decrease rapidly, and practically completely if the depth of the water table is less than 20cm below the surface.

Equation 3.6 includes the essential elements for estimating N₂O emissions from re-wetted peatlands or organic soils:

EQUATION 3.6

$$N_2O_{rewetted} = N_2O_{soil} + N_2O_{biomass\ burn} + N_2O_{soil\ burn}$$

407 Where:

408 $N_2O_{rewetted}$ = N₂O emissions from re-wetted peatlands or organic soils, tonnes N₂O yr⁻¹

409 N_2O_{soil} = N₂O emissions from the soil pool of re-wetted peatland or organic soils, tonnes C yr⁻¹

410 $N_2O_{biomass\ burn}$ = N₂O emissions from biomass burning on rewetted peatlands or organic soils, tonnes
411 N₂O yr⁻¹

412 $N_2O_{soil\ burn}$ = N₂O emissions from soil or peat burning on rewetted peatlands or organic soils, tonnes N₂O
413 yr⁻¹

414 This chapter only covers N₂O_{soil}. Generic methodologies for estimating N₂O emissions from the burning of
415 vegetation and dead organic matter are provided in chapter 2 of volume 4 in the existing 2006 GLs, while
416 methodologies specific to vegetation and DOM burning in forest land, cropland, grassland and wetlands are
417 provided in chapters 3-6 of volume 4. When burning occurs, emissions from the burning of soil should be
418 estimated using country-specific (Tier 2) emission factors.

419 Under Tier 1, emissions of nitrous oxides from re-wetted soils are assumed to be insignificant¹.

420 At higher tiers, countries are encouraged to examine the pattern of N₂O emissions from re-wetted peatlands and
421 organic soils by climate zone and peat type, in particularly tropical climate. Country specific emission factors
422 should take into account fluctuations of the water table depth, which controls oxygen availability for nitrification,
423 and previous land use, which may have resulted in top soil enrichment (Nagata et al., 2006; 2010). Caution
424 should be exerted to avoid double-counting N₂O emissions from re-wetted sites that can be directly linked to
425 allochthonous (external) nitrogen input, as these emissions may already be reported as indirect emissions from
426 anthropogenic N input within the watershed.

427

428 3.3 COMPLETENESS, TIME SERIES, 429 CONSISTENCY, AND QA/QC

430 3.3.1 Completeness

431 Complete greenhouse gas inventories will include estimates of emissions from all greenhouse gases emissions
432 and removals on rewetted and restored peatlands for which default guidance is provided in this chapter, for all
433 peatland types or organic soils that occur on the national territory.

434 It is good practice to ensure that all *drained* peatlands and organic soils on the national territory are already
435 included in the national inventory, prior to estimating emissions and removals from rewetted peatlands or
436 organic soils.

437 As for all land categories, countries are encouraged to monitor the fate of rewetted and restored peatlands, and
438 avoid double counting emissions reported from lands in various categories. The greenhouse gas balance of re-
439 wetted lands with peat or organic soils should include all applicable carbon pools, while avoiding double-
440 counting between carbon pools, especially if flux-based estimation methodologies are combined with stock-
441 change approaches.

442 3.3.2 Developing a consistent time series

443 General guidance on ensuring time series consistency can be found in Chapter 7 of this Supplement. The
444 emission and removal estimation method should be applied consistently to every year in the time series, at the
445 same level of spatial disaggregation. Moreover, when country-specific data are used, national inventory agencies
446 should use the same measurement protocol (sampling strategy, method, etc.) throughout the time series, or use
447 documented, unbiased methods to ensure the representativeness and applicability of emission factors.

¹ Augustin, 2003; Augustin and Chojnicki, 2008; Drösler, 2005; Hendriks et al., 2005; Jungkunst and Fiedler, 2007; Wild et al., 2001, Wilson et al., in press.

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448 **3.3.3 Quality Assurance and Quality Control (QA/QC)**

449 Quality assurance/quality control (QA/QC) procedures should be developed and implemented as outlined in
450 Chapter 7 of this Supplement. It is good practice to develop additional, category-specific quality control and
451 quality assurance activities if higher tier methods are used to quantify emissions and removals from this category.
452 Where country-specific emission factors are being used, they should be based on high quality experimental data,
453 developed using a rigorous measurement programme, and be adequately documented, preferably in the peer-
454 reviewed, scientific literature.

455 **3.3.4 Reporting and Documentation**

456 **EMISSION FACTORS**

457 The scientific basis of country-specific emission factors, parameters and models should be fully described and
458 documented. This includes describing the input parameters, the derivation of emission factors, parameters and
459 models and the sources of uncertainty, as well as justifying their representativeness.

460 It is good practice to clearly document the carbon pools that are captured by flux-based emission factors and the
461 completeness in terms of pool coverage.

462

463 **ACTIVITY DATA**

464 Sources of all activity data used in the calculations (data sources, databases and soil map references) should be
465 recorded, plus (subject to any confidentiality considerations) communication with industry. This documentation
466 should cover the frequency of data collection and estimation, and estimates of accuracy and precision, and
467 reasons for significant changes in activity data.

468 Information should be provided, for each land-use category, on the proportion of drained and re-wetted areas
469 with organic soils. Overall, the sum of re-wetted areas with peat or organic soils reported under each land use
470 categories should equal the total national area of re-wetted peatlands or organic soils.

471

472 **TREND ANALYSIS**

473 Significant fluctuations in emissions and removals between years should be explained. A distinction should be
474 made between changes in activity levels and changes in emission factors, parameters and methods from year to
475 year, and the reasons for these changes documented. If different emission factors, parameters and methods are
476 used for different years, the reasons for this should be explained and documented.

477

478 **References**

- 479 Alm J., Talanov A., Saarnio S., Silvola J., Ikkonen E., Aaltonen H., Nykänen H., Martikainen P.J. 1997.
480 Reconstruction of the carbon balance for microsites in a boreal oligotrophic pine fen, Finland. *Oecologia*
481 110: 423-431.
- 482 Augustin, J., Merbach, W., Käding, H., Schmidt, W. & Schalitz, G. 1996. Lachgas- und Methanemission aus
483 degradierten Niedermoorstandorten Nordostdeutschlands unter dem Einfluß unterschiedlicher
484 Bewirtschaftung. In: Alfred-Wegener-Stiftun (ed.) Von den Ressourcen zum Recycling. Berlin, Ernst &
485 Sohn. pp 131-139.
- 486 Augustin, J. & Merbach, W. 1998. Greenhouse gas emissions from fen mires in Northern Germany:
487 quantification and regulation. In: Merbach, W. & Wittenmayer, L. Beiträge aus der Hallenser
488 Pflanzenernährungsforschung, pp. 97-110
- 489 Augustin, J. 2003. Gaseous emissions from constructed wetlands and (re)flooded meadows. *Publicationes*
490 *Instituti Geographici Universitatis Tartuensis* 94: 3-8
- 491 Augustin, J. & B. Chojnicki, 2008. Austausch von klimarelevanten Spurengasen, Klimawirkung und
492 Kohlenstoffdynamik in den ersten Jahren nach der Wiedervernässung von degradiertem
493 Niedermoorgrünland. In Gelbrecht, J., D. Zak & J. Augustin (eds) Phosphor- und Kohlenstoff-Dynamik und
494 Vegetationsentwicklung in wiedervernässten Mooren des Peenetales in Mecklenburg-Vorpommern – Status,
495 Steuergrößen und Handlungsmöglichkeiten. Berichte des IGB Heft 26. IGB, Berlin: 50–67.
- 496 Augustin, unpubl., cited in Couwenberg, J., Thiele, A., Tanneberger, F., Augustin, J., Bärtsch, S., Dubovik, D.,
497 Liashchynskaya, N., Michaelis, D., Minke, M., Skuratovich, A. & Joosten, H. (2011) Assessing greenhouse
498 gas emissions from peatlands using vegetation as a proxy. *Hydrobiologia*, 674, 67–89.
- 499 Billett M. F., Palmer M., Hope D., Deacon C., Storeton-West R., Hargreaves K. J., Flechard C. & Fowler D.
500 2004. Linking land-atmosphere-stream carbon fluxes in a lowland peatland system. *Global Biogeochemical*
501 *Cycles* 18(GB1024): doi:10.1029/2003GB002058.
- 502 Billett M. F. & Moore T. R. 2008. Supersaturation and evasion of CO₂ and CH₄ in surface waters at Mer Bleue
503 peatland, Canada. *Hydrological Processes* 22: 2044-2054.
- 504 Billett M. F., Charman C., Clark I., Evans C. D., Evans M. G., Ostle N., Worrall F., Burden A., Dinsmore K. J.,
505 Jones T., McNamara N. P., Parry L., Rowson J. G. & Rose R. 2010. Carbon balance of UK peatlands: current
506 state of knowledge and future research challenges. *Climate Research* 45: 13-20.
- 507 Bortoluzzi, E., Epron, D., Siegenthaler, A., Gilbert, A. & Butler, A. 2006. Carbon balance of a European
508 mountain bog at contrasting stages of regeneration. *New Phytologist* 172: 708-718
- 509 Bubier J.L., Moore T.R., Roulet N.T. 1993. Methane emissions from wetlands in the midboreal region of
510 Northern Ontario, Canada. *Ecology* 74(8): 2240-2254.
- 511 Cleary J, Roulet NT, Moore TR. 2005. Greenhouse gas emissions from Canadian peat extraction, 1990-2000: A
512 life-cycle analysis, *Ambio*, 34, 456-461.
- 513 Clymo R.S., Reddaway E.J.F. 1971. Productivity of Sphagnum (Bog-moss) and peat accumulation.
514 *Hydrobiologia* 12: 181-192. (cited in: Bartlett K.B. & Harris R.C. 1993. Review and assessment of Methane
515 Emissions from Wetlands. *Chemosphere*. Vol.26, Nos. 1-4: 261-320.)
- 516 Couwenberg, J., Thiele, A., Tanneberger, F., Augustin, J., Bärtsch, S., Dubovik, D., Liashchynskaya, N.,
517 Michaelis, D., Minke, M., Skuratovich, A. & Joosten, H. (2011) Assessing greenhouse gas emissions from
518 peatlands using vegetation as a proxy. *Hydrobiologia*, 674, 67–89.
- 519 Crill, unpublished data. (cited in: Bartlett K.B. & Harris R.C. 1993. Review and assessment of Methane
520 Emissions from Wetlands. *Chemosphere*. Vol.26, Nos. 1-4: 261-320.)
- 521 Dinsmore K. J., Billett M. F., Skiba U. M., Rees R. M., Drewer J. & Helfter C. 2010. Role of the aquatic pathway
522 in the carbon and greenhouse gas budgets of a peatland catchment. *Global Change Biology* 16: 2750-2762.
- 523 Dise N.B., Gorham E. 1993. Environmental Factors Controlling Methane Emissions from Peatlands in northern
524 Minnesota. *Journal of Geophysical Research* 98 Nr. D6: 10583-10594.
- 525 Drewer J., Lohila A., Aurela M., Laurila T., Minkkinen K., Penttilä T., Dinsmore K. J., McKenzie R. M., Helfter
526 C., Flechard C., Sutton M. A. & Skiba U. M. 2010. Comparison of greenhouse gas fluxes and nitrogen
527 budgets from an ombrotrophic bog in Scotland and a minerotrophic sedge fen in Finland. *European Journal of*
528 *Soil Science*: 10.1111/j.1365-2389.2010.01267.x

First-order Draft

- 529 Drösler, M. 2005. Trace gas exchange and climatic relevance of bog ecosystems, southern Germany. PhD thesis,
530 Technische Universität München. 182p.
- 531 Flessa, H., Wild, U., Klemisch, M. & Pfadenhauer, J. 1997. C- und N-Stoffflüsse auf
532 Torfstichsimulationsflächen im Donaumoos. Zeitschrift für Kulturtechnik und Landentwicklung 38:11-17.
- 533 Gauci V., Dise N. 2002. Controls on suppression of methane flux from a peat bog subjected to simulated acid rain
534 sulfate deposition. *Global Biogeochemical Cycles* 16 Nr. 1: 4-1 to 4-12.
- 535 Glatzel, S., Koebisch, F., Beetz, S., Hahn, J., Richter, P., Jurasinski, G., 2011. Maßnahmen zur Minderung der
536 Treibhausgasfreisetzung aus Mooren im Mittleren Mecklenburg. *Telma*. 4: 85-106.
- 537 Harriss R.C., Sebacher D.I., Day F.P. 1982. Methane flux in the Great Dismal Swamp. *Nature* 297: 673-674.
538 (cited in: Bartlett K.B. & Harris R.C. 1993. Review and assessment of Methane Emissions from Wetlands.
539 *Chemosphere*. Vol.26, Nos. 1-4: 261-320.)
- 540 Hendriks, D.M.D., van Huissteden, J., Dolma, A.J. & van der Molen, M.K. 2007. The full greenhouse gas
541 balance of an abandoned peat meadow. *Biogeosciences* 4: 411-424
- 542 Jungkunst, H.F. & Fiedler, S. 2007. Latitudinal differentiated water table control of carbon dioxide, methane and
543 nitrous oxide fluxes from hydromorphic soils: feedbacks to climate change. *Global Change Biology* 13:
544 2668-2683
- 545 Juottonen, H., Hynninen, A., Nieminen, M., Tuomivirta, T., Tuittila, E-S., Nousiainen, H., Yrjälä, K.,
546 Tervahauta A., Fritze, H. Methane-cycling microbial communities and methane emission in natural and
547 restored peatland buffer areas. Submitted to *Applied and Environmental Microbiology* in January 2012.
- 548 Koehler A-K, Sottocornola M, Kiely G, 2011. How strong is the current carbon sequestration of an Atlantic
549 blanket bog? *Global Change Biology*, 17, 309–319.
- 550 Komulainen, V.-M., H. Nykanen, P. J. Martikainen, Laine, J. 1998. Short-term effect of restoration on vegetation
551 change and methane emissions from peatlands drained for forestry in southern Finland. *Canadian Journal of*
552 *Forest Research* 28: 402-411.
- 553 Laine J., Silvola J., Tolonen K., Alm J., Nykänen H., Vasander H., Sallantausta T., Savolainen I., Sinisalo J.,
554 Martikainen P.J. 1996. Effect of Water-level Drawdown on Global Climatic Warming: Northern Peatlands.
555 *Ambio* Vol. 25 No. 3: 179-184. Royal Swedish Academy of Sciences.
- 556 Nagata, O., Takakai, F. & Hatano, R. 2005. Effect of Sasa invasion on global warming potential in Sphagnum
557 dominated poor fen in Bibai, *Phyton*, 45:299-307.
- 558 Nagata, O., Yazaki, T., Yanai, Y. 2010. Nitrous oxide emissions from drained and mineral soil-dressed peatland
559 in central Hokkaido, Japan. *Journal of Agricultural Meteorology*, 66:23-30
- 560 Nilsson M, Sagerfors J, Buffam I, Laudon H, Eriksson T, Grelle A, Klemedtsson L, Weslien P, Lindroth A,
561 2008. Contemporary carbon accumulation in a boreal oligotrophic minerogenic mire – a significant sink after
562 accounting for all C-fluxes *Global Change Biology* 14, 2317–2332.
- 563 Nykänen H., Alm J., Lang K., Silvola J., Martikainen P. 1995. Emissions of CH₄, N₂O and CO₂ from a virgin
564 fen and a fen drained for grassland in Finland. *Journal of Biogeography* 22: 351-357.
- 565 Saarnio S, Morero M, Shurpali NJ, Tuittila E-S, Mäkilä M, Alm J (2007) Annual CO₂ and CH₄ fluxes of
566 pristine boreal mires as a background for the lifecycle analyses of peat energy, *Boreal Environment Research*
567 12: 101-113.
- 568 Scottish Executive. 2007. *Ecosse - Estimating Carbon in Organic Soils, Sequestration and emissions*. Scottish
569 Executive, Edinburgh. <http://www.scotland.gov.uk/Publications/2007/03/16170508> [febr. 2008]. 177 p.
- 570 Shannon R.D., White J.R. 1994. A three-year study of controls on methane emissions from two Michigan
571 peatlands. *Biogeochemistry* 27: 35-60.
- 572 Sommer, M., Fiedler, S., Glatzel, S. & Kleber, Markus. 2003. First estimates of regional (Allgäu, Germany) and
573 global CH₄ fluxes from wet colluvial margins of closed depressions in glacial drift areas. *Agriculture*
574 *Ecosystems & Environment* 103: 251-257
- 575 Strack M., Toth K., Bourbonniere R. A. & Waddington J. A. 2011. Dissolved organic carbon production and
576 runoff quality following peatland extraction and restoration. *Ecological Engineering* 37: 1998-2008.
- 577 Tauchnitz, N., Brumme, R., Bernsdorf, S. & Meissner, R. 2008. Nitrous oxide and methane fluxes of a pristine
578 slope mire in the German National Park Harz Mountains. *Plant and Soil* 303, 131-138

- 579 Tuittila E.-S., Komulainen V.-M., Vasander H., Nykänen H., Martikainen P.J., Laine J. 2000. Methane dynamics
580 of a restored cut-away peatland. *Global Change Biology* 6: 569-581.
- 581 Urbanová, Z., Pícek, T., Hájek, T., Bufková, I., Tuittila, E.-S. Impact of drainage and restoration on vegetation
582 and carbon gas dynamics in Central European peatlands. Conditionally accepted to *Plant Ecology and*
583 *Diversity*, 2012.
- 584 Verma S.B., Ullman F.G., Billesbach D., Clement R.J., Kim J., Verry E.S. 1992. Eddy correlation measurements
585 of methane flux in a northern peatland ecosystem. *Bound. Layer Meteorol.* 58:289-304. (cited in: Bartlett
586 K.B. & Harris R.C. 1993. Review and assessment of Methane Emissions from Wetlands. *Chemosphere.*
587 Vol.26, Nos. 1-4: 261-320.)
- 588 Von Arnold, K. 2004. Forests and greenhouse gases - fluxes of CO₂, CH₄ and N₂O from drained forests on
589 organic soils. *Linköping Studies in Arts and Science* no 302. 48p.
- 590 Waddington J.M., Price J.S. 2000. Effect of peatland drainage, harvesting, and restoration on atmospheric water
591 and carbon exchange. *Physical Geography* 21, 5: 433-451.
- 592 Waddington J.M., Roulet N.T. 2000. Carbon balance of a boreal patterned peatland. *Global Change Biology* 6:
593 87-97.
- 594 Whiting G.J. & Chanton J.P. 2001. Greenhouse carbon balance of wetlands: methane emission versus carbon
595 sequestration. *Tellus* 53B: 521-528.
- 596 Wickland K.P., Striegl R.G., Mast M.A., Clow D.W. 2001. Carbon gas exchange at a southern Rocky Mountain
597 wetland 1996-1998. *Global Biogeochemical Cycles* 15 Nr. 2: 321-335.
- 598 Wild, U., Kamp, T., Lenz, A., Heinz, S. & Pfadenhauer, J. 2001. Cultivation of *Typha* spp. in constructed
599 wetlands for peatland restoration. *Ecological Engineering* 17:49-54
- 600 Wilson D., Alm J., Laine J., Byrne K. A., Farrell E. P. & Tuittila E.-S. 2009. Rewetting of cutaway peatlands:
601 Are we re-creating hotspots of methane emissions? *Restoration Ecology* 17(6): 796-806 doi: 10.1111/j.1526-
602 100x.2008.00416.x.
- 603 Wilson D., Renou-Wilson F., Farrell C., Bullock C. & Müller C. In press. Carbon Restore - The potential of
604 peatlands for carbon sequestration. *Climate Change Research Programme Report*. Environmental Protection
605 Agency. Johnstown Castle, Co. Wexford, Ireland.
- 606 Wilson J.O., Crill P.M., Bartlett K.B., Sebacher D.I., Harriss R.C., Sass R.L. 1989. Seasonal variation of
607 methane emissions from a temperate swamp. *Biogeochem.* 8: 55-71. (cited in: Bartlett K.B. & Harris R.C.
608 1993. Review and assessment of Methane Emissions from Wetlands. *Chemosphere.* Vol.26, Nos. 1-4: 261-
609 320.)
- 610 Worrall F., Burt T. P., Rowson J. G., Warburton J. & Adamson J. K. 2009. The multi-annual carbon budget of a
611 peat-covered catchment. *Science of the Total Environment* 407: 4084-4094
- 612 Yli-Petäys, M., Laine, J., Vasander, H. & Tuittila, E.-S. 2007: Carbon gas exchange of a re-vegetated cut-away
613 peatland five decades after abandonment. *Boreal Environmental Research* 12: 177-190.
- 614
- 615

616 **Annex 3A.1 Estimation of default emission factors for CO₂-C in** 617 **rewetted peatlands and organic soils**

618 An extensive literature review was conducted to collate all greenhouse gas (GHG) studies that are currently
619 available for natural/undrained and rewetted peatlands and organic soils. Literature sources included both
620 published and non-published (grey literature) studies. In the case of the latter, expert judgement was exercised as
621 to whether the study was scientifically acceptable for inclusion. A greater amount of data from natural
622 (undrained) is available than from re-wetted peatlands. A detailed database of annual CO₂ fluxes was then
623 constructed to determine the main drivers (if any) of CO₂ dynamics in rewetted peatlands and organic soils.
624 When available, the following parameters were extracted and included in the database for analysis: climate zone
625 (latitude), peatland types, mean annual water table, median annual water table (as well as minimum and
626 maximum), soil pH, peat thickness, peat C/N ratio, degree of humification, soil moisture, soil bulk density, plant
627 cover and species, previous land-use and time since re-wetting.

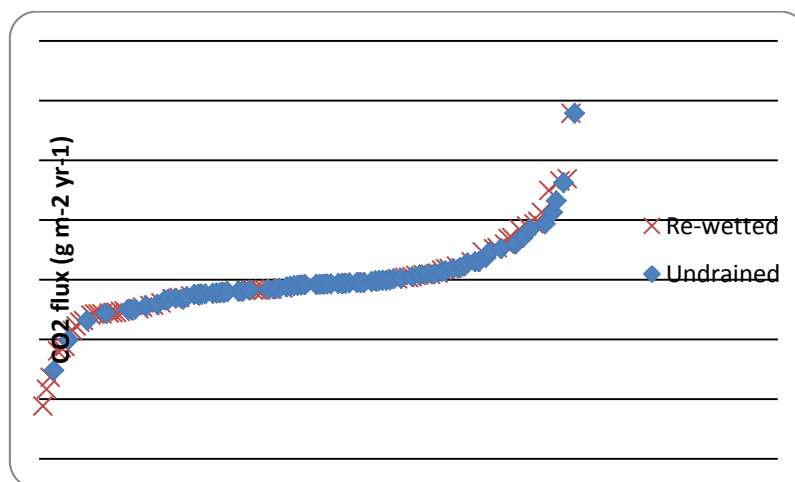
628 To determine Tier 1 CO₂ emission factors, descriptive statistics allowed the data to be grouped by (1) *climate*
629 *region*, (2) *peatland type (bog or fen)* and (3) *climate region and peatland type*, and descriptive analysis for each
630 group was computed. Due to difficulties in up-scaling flux data or incomplete data entries a proportion of the
631 CO₂ flux entries in the database (12-33%) were not used in the analysis. Studies that reported daily CO₂ flux
632 values in the database were not used. Seasonal CO₂ fluxes (typically May to October) were converted to annual
633 fluxes by assuming that 15% of the flux occurs in the non-growing season (Saarnio et al. 2007). Tropical
634 peatlands were initially included in the database and although several studies were first extracted, the number of
635 usable entries was deemed very limited (one re-wetted and three undrained/natural entries) and were therefore
636 excluded thereafter.

637 While all re-wetted entries were included (n=62), data from natural/undrained were further extracted and only
638 entries (n=82) from sites with a mean annual water table of no less than 30 cm was used as a proxy for re-wetted
639 sites. These entries were well distributed between bogs (n=79) and fens (n= 65). There were 63 boreal entries
640 compared with 81 temperate entries with temperate bogs being the most common studied type.

641 While a total 142 fluxes were entered for the analysis (Figure 3A.1), 130 entries were used in the final database
642 to calculate the emission factors.

643 Firstly, a comparison was made between re-wetted sites and natural/undrained sites. The range of fluxes from
644 both groups is analogous (Figure 3A.1).

645 **Figure 3A.1**



646 The fitted regression lines ($\text{CO}_2 \text{ flux} = a + b1 * \text{WT}$) were compared between re-wetted and natural/undrained
647 groups for each category. The two groups were treated as being non-significantly different when it was
648 ascertained statistically that $b1 \pm \text{S.E. (rewetted)}$ fitted within $b1 - \text{S.E.}$ and $b1 + \text{S.E.}$ for the natural/undrained
649 group. This was the case for all but the temperate fen groups. Therefore, emission factors were calculated using
650 re-wetted and natural/undrained entries for all categories except for temperate fens where only re-wetted entries
651 (n=17) were used in the calculation of the emission factor. Total number of entries used overall was 130.
652

653

654 **Annex 3A.2 Estimation of default emission factors for CO₂-DOC**
 655 **in rewetted peatlands and organic soils**

656 Collated data from seven peat re-wetting studies suggest a median DOC reduction of 36%, with a range of 1-
 657 69%. While the number of studies is limited, and results are fairly variable, the median reduction is almost
 658 exactly equivalent to the observed increase following drainage (a 33% decrease would be required to fully
 659 reverse a 50% increase). A study by Glatzel et al. (2003) observed similar DOC concentrations in re-wetted and
 660 restored bogs (previously used for peat harvesting) as in a nearby intact reference bog. Therefore, there is some
 661 evidence to suggest that peat re-wetting will return DOC loss fluxes to natural levels. This is likely to occur after
 662 an initial pulse of DOC associated with disturbance during the re-wetting process (e.g. Zak and Gilbrecht, 2007;
 663 Worrall et al., 2007, Strack et al., 2011), which may contribute to the variability shown in Table 3A.X, where
 664 some measurements were made soon after re-wetting.

665 The values derived for the estimation of DOC_{FLUX_NATURAL} are explained in Annex X of Chapter 2 as well as an
 666 explanation of the estimated F_{DOC-CO₂}

Previous land-use	Country	Study	DOC (mg/l)		DOC _{RE-WET} (%)
			Drained	Re-wetted	
Peat extraction bog	Canada	Glatzel et al (2003)	110	70	-36%
Drained blanket bog	UK	Wallage et al (2006)	43	13	-69%
Drained blanket bog	UK	Armstrong et al (2010)	34	30	-10%
Drained blanket bog	UK	Gibson et al (2009)	39	39	-1%
Drained agricultural fen	Germany	Höll et al (2009)	86	57	-34%
Previous land-use	Country	Study	DOC (g C m ⁻² y ⁻¹)		DOC _{RE-WET} (%)
			Undrained	Drained	
Peat extraction bog	Canada	Waddington et al (2008)	7.5	3.5	-53%
Drained blanket bog	UK	O'Brien et al (2008)	7.0	4.1	-41%

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669 **Annex 3A.3 Estimation of default emission factors for CH₄-C in** 670 **rewetted peatlands and organic soils**

671 The same literature database and general approach were used to develop default CH₄ emission factors as was
672 described in Annex 3.1. Similarly, a greater amount of data from natural (undrained) peatlands is available than
673 from re-wetted peatlands. A detailed database of annual CH₄ fluxes was then constructed to determine the main
674 drivers (if any) of CH₄ emissions in rewetted peatlands and organic soils. Where possible, the analysis
675 considered the same parameters : climate zone (latitude), peatland types, mean annual water table, median
676 annual water table (as well as minimum and maximum), soil pH, peat thickness, peat C/N ratio, degree of
677 humification, soil moisture, soil bulk density, plant cover and species, previous land-use and time since re-
678 wetting.

679 Methane fluxes from rewetted boreal peatlands (mean 80.3 kg CH₄-C ha⁻¹ yr⁻¹; variance; n=27²) are
680 significantly lower than from undrained (pristine) sites (mean 205.4 kg CH₄-C ha⁻¹ yr⁻¹; variance; n=87³). The
681 increase in efflux with rising water level is similar, but significantly steeper for undrained sites (n=68 pairs)
682 compared to rewetted sites (n= 13 pairs). Methane efflux from rewetted fens (mean 161.6 kg CH₄-C ha⁻¹ yr⁻¹;
683 variance; n=12) is an order of magnitude higher than efflux from rewetted bogs (mean 16.2 kg CH₄-C ha⁻¹ yr⁻¹;
684 variance; n=12), which is mirrored by efflux values from undrained fens (mean 392.5 kg CH₄-C ha⁻¹ yr⁻¹;
685 variance; n=33) and bogs (65.6 kg CH₄-C ha⁻¹ yr⁻¹; variance; n=42).

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687 Whereas methane fluxes from rewetted temperate peatlands (mean 416.1 kg CH₄-C ha⁻¹ yr⁻¹; variance; n=47⁴)
688 are considerably higher than from undrained peatlands (mean 148.7 kg CH₄-C ha⁻¹ yr⁻¹; variance; n=61⁵), this
689 finding is based mainly on a small number of extremely high efflux values from sites on enriched agricultural
690 soil that were flooded during rewetting (Augustin & Chojnicki 2008; Glatzel et al., 2011). The increase in efflux
691 with rising water level is not significantly different between undrained (n=39 pairs) and rewetted sites (n=40
692 pairs), nor between undrained bogs (n=35) and undrained fens (n=22). Flux values from rewetted sites do show
693 higher variance, however, which cannot be explained by available parameters (peat type, water level dynamics,
694 pH, C/N ratio, vegetation), although the highest values were measured in recently rewetted fen peatlands,
695 formerly used as high-intensity grassland (time since rewetting ≤ 2 years). Methane effluxes from rewetted
696 temperate bogs are an order of magnitude lower than from fens, but measurements are restricted to only three
697 bog sites. Combined, the increase in efflux with rising water level in undrained and rewetted sites does not show
698 a significant difference between bogs (n=30 pairs) and fens (n=49 pairs). The emission factors presented are
699 based on the total dataset of rewetted and undrained bogs (n=41) and fens (n=63). Because bogs have more
700 relatively dry microsites and the dataset for fens includes the extreme values mentioned above, the EF for
701 temperate bogs is lower than for fens.

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704 Very few methane flux data exist on rewetted tropical peatlands. Comparison of hourly flux data suggests that
705 efflux in SE Asian peat swamps (the largest extent of tropical peatland) will be much lower than in
706 temperate/boreal regions (Couwenberg et al., 2010). Efflux from rice padi on peat soil is comparable to current
707 IPCC estimates (Couwenberg 2011). Efflux from Papyrus marshes has been measured, paper is underway...

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² Juottonen et al., 2012; Komulainen et al., 1998; Tuittila et al., 2000 ; Urbanová et al., 2012 ; Yli-Petäys et al., 2007

³ Alm et al., 1997; Bubier et al., 1993; Clymo & Reddaway, 1971; Drewer et al., 2010; Gauci & Dise 2002; Laine et al., 1996 ; Nykänen et al., 1995 ; Verma et al., 1992 ; Waddington & Roulet 2000 ; Whiting & Chanton 2001

⁴ Augustin & Chojnicki 2008; Augustin & Merbach 1998; Augustin 2003; Augustin in Couwenberg et al. 2011; Cleary et al., 2005; Drösler 2005; Flessa et al., 1997; Glatzel et al., 2011; Hendriks et al., 2007; Jungkunst & Fiedler 2007; Waddington & Price 2000; Wild et al., 2001; Wilson et al., 2009; Wilson et al., in press

⁵ Augustin & Merbach 1998; Augustin 2003; Augustin et al., 1996; Augustin in Couwenberg et al. 2011; Bortoluzzi et al. 2006; Crill in Bartlett & Harris 1993; Dise & Gorham 1993; Drösler 2005; Harriss et al., 1982; Koehler et al., 2010; Nagata et al., 2005; Nilsson et al., 2008; Roulet et al., 2007; Scottish Executive, 2007; Shannon & White 1994; Sommer et al., 2003; Tauchnitz et al., 2008; Von Arnold 2004; Waddington & Price 2000; Wickland et al., 2001; Wilson et al., 1989

710 **APPENDIX 3.1 CO₂ Emissions/Removals from Re-Wetted**
711 **Peatlands or Organic Soils in Tropical Climates: a Basis for**
712 **Future Methodological Development**

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717 **APPENDIX 3.1 Time-Series of EF_{DOC Rewetted}; a Basis For**
718 **Future Methodological Development**

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